Advances on Transition Metal Oxides Catalysts for Formaldehyde Oxidation: A

Review

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1 Abstract

This review highlights recent advances in the development of transition metal based catalysts for 2 formaldehyde oxidation, particularly the enhancement of their catalytic activity for low 3 temperature oxidation. Various factors that enhance low temperature activity are reviewed, such 4 as morphology and tunnel structures, synthesis methods, specific surface area, amount and type of 5 6 active surface oxygen species, oxidation state and density of active sites are discussed. In addition, catalyst immobilization for practical air purification, reaction mechanism of formaldehyde 7 oxidation and the reaction parameters affecting the overall efficiency of the reaction are also 8 reviewed. 9

10 Keywords: formaldehyde, transition metal oxides, catalysts, oxidation

11 **1. Introduction**

Formaldehvde (HCHO) is one of the main sources of hazardous indoor air pollution. Furniture and 12 building materials such as composite wood, particle board, vinyl coverings and adhesives are some 13 of the major indoor sources of HCHO emissions (1,2). A comprehensive review of both the indoor 14 and outdoor sources of HCHO and its concentrations in various indoor environments can be found 15 in the work of Salthammer et al. (3). Exposure to HCHO may cause adverse health effects on 16 humans such as irritation to eyes, nose and throat, headache, fatigue, edema, severe allergic 17 reaction and dermatitis (4,5). In 2006, the International Agency for Research on Cancer (IARC) 18 19 under the World Health Organization (WHO) classified HCHO as a carcinogen to humans. There are sufficient evidences to ascertain that it causes nasopharyngeal cancer and there are also strong 20 21 indications that it may also cause sinonasal cancer and leukemia over long term exposure (6). As 22 such there are various international guidelines and recommended thresholds for indoor air HCHO concentrations, a summary of which can be found in the review of Salthammer et al. (3). The WHO 23 proposes a short term HCHO exposure limit (30 minutes) of 0.1 mg/m³ for the avoidance of 24 sensory irritation and a long term exposure limit of 0.2 mg/m³ for protection against long term 25 health effects (7). Thus, the effective removal of HCHO from indoor air is imperative to improving 26 indoor air quality and safeguarding human health. 27

Various techniques for HCHO removal have been investigated and reported in the literature, including adsorption (8-12), photo-catalytic oxidation (13-17) and catalytic oxidation (thermal and non-thermal). The effectiveness of physical adsorption of HCHO on adsorbents such as activated carbon (AC) is constrained by the material's maximum adsorption capacity, relative humidity (RH) or moisture deactivation and also by the environmental risk when desorption occurs during regeneration (18,19). Photo-catalytic oxidation using ultraviolent light on the other hand may lead to the generation of toxic by-products (*20*). Catalytic oxidation is able to achieve complete conversion of HCHO to H₂O and CO₂ without the formation of harmful by-products or secondary pollutants (*21*). This could even be achieved at room temperature especially with noble metal catalysts (*22-26*), thus making it the most promising HCHO removal technique (*18*).

Noble metal-based catalysts: Pt, Au, Pd, Rh, Ru and Ag supported on conventional materials such 38 39 as TiO₂, SiO₂, Al₂O₃ and zeolites (27-35); single transition metals (36-47) and transition metalbased composites (48-51) exhibit excellent HCHO oxidation activities at temperatures lower than 40 100°C and even at room temperature. Comprehensive reviews on noble metal catalysts for HCHO 41 oxidation have been conducted (4, 25), whilst here a summary of some of the noble metal catalysts 42 and their reaction conditions are presented in Table 1. However, the industrial applications of noble 43 metal catalysts are restricted by their high costs, limited resources and poor thermal stabilities (53). 44 Hence recent research efforts have been focused on the development of relatively cheap materials 45 for low temperature HCHO catalytic oxidation (54-56). More abundant and cost effective metal 46 47 oxides especially those of the transition metals, including single transition metals and transition metal based composites are shown to be active for HCHO oxidation. However their relatively low 48 activity compared to noble metals catalysts has motivated effort for further improvement. 49 50 Therefore, the development of highly active and cost effective catalysts for HCHO oxidation is still a major challenge for practical application. 51

A number of review papers have been published on HCHO removal. Pei and Zhang (*52*) reviewed chemisorption method and catalytic oxidation of HCHO majorly on noble metal catalysts. The photo-catalytic oxidation of HCHO using TiO₂ as a photo-catalyst both in aqueous and gaseous mediums and the effects of different light sources and photoreactors have also been reviewed in a short article (*57*). More recently, Bai et al. (*4*) reviewed the catalysts for HCHO oxidation with a 57 major focus on the factors affecting the activity of noble metal catalysts. From the literature review, 58 it can be seen that significant progress has been made on improving the activity of transition metal 59 based catalysts and their immobilization for practical application since 2000. However, to the best 60 of our knowledge, the development of transition metal based catalysts for HCHO degradation has 61 not been reviewed in any detail so far. Therefore, we consider this review fills a key gap for this 62 active area of heterogeneous catalysis.

63 This review focuses exclusively and extensively on recent developments over the past one and half decade towards enhancing the activity of transition metal catalysts for low temperature HCHO 64 oxidation, considering their cost reduction potential, activity and stability. Various factors that 65 enhance their catalytic activities are discussed, including preparation methods, morphology and 66 structure, specific surface area, concentration of surface active oxygen species, oxygen mobility 67 and metal active sites. The influence of reaction parameters such as relative humidity, HCHO 68 concentration and space velocity are also reviewed. The reaction mechanisms of HCHO oxidation 69 on transition metal oxides and their immobilization on suitable substrate materials for application 70 in air purification are also elaborated and finally areas for further investigations to achieve higher 71 activities at low temperatures are proposed. 72

73 2. Transition metal based catalysts and their performances

As mentioned earlier, recent research efforts have been focused on the utilization and efficiency improvement of transition metal based catalysts for low temperature oxidation of HCHO. Transition metal oxides are relatively cheap, abundant and have also been presented to be active for HCHO oxidation at low temperatures (*21,67,68*). Different terminologies such as HCHO removal, conversion, degradation and elimination are used in the literature to refer to the efficiency of HCHO oxidation. Hereinafter, efficiency of HCHO oxidation is referred to as conversion where oxidation efficiency was reported as a function of CO₂ generation/concentration in the effluent
stream and as removal where efficiency was reported based on residual HCHO concentration in
the effluent gas stream (see footnotes of Table 1&2).

83 Some conventional transition metal-based catalysts have been reported to show good catalytic activities for HCHO oxidation. Sekine (2) first demonstrated the catalytic oxidation of HCHO over 84 85 metal oxide catalysts. Of the investigated catalysts: CoO, MnO₂, TiO₂, CeO₂ and Mn₃O₄, MnO₂ was described to exhibit the highest catalytic activity. Metal oxides in the conventional bulk form 86 possess low catalytic activity for HCHO oxidation owing to inferior surface properties (69). On 87 88 the contrary, specially synthesized nanostructured transition metal catalysts of similar compositions to their bulk counterparts exhibit improved morphologies and surface properties and 89 hence higher activities for HCHO oxidation (21,70). As such, current investigations on transition 90 metal-based catalysts are focused on improving and developing materials with enhanced 91 morphologies and structures. The structure of these catalysts and their morphologies improve their 92 93 activity through the improvement of parameters (discussed in Section 4) such as specific surface area, high surface reducibility, porosity, active surface species, active lattice and 94 surface adsorbed oxygen species and tunnel size and structure (54,55). Transition metal-based catalysts 95 96 for HCHO oxidation found in the literature could basically be classified as single/mono-metal and composite metal oxides. A summary of both group of catalysts and their activities and conditions 97 of reaction are respectively presented in Table 2 and 3. 98

99 2.1 Single transition metal based catalysts

100 2.1.1 Manganese oxide based catalysts

101 Manganese oxide is the most widely explored transition metal catalyst for HCHO oxidation owing

to its high catalytic activity, thermal stability, existence in various crystal morphologies such as α -

, β -, γ - and δ -MnO_x (71) and several tunnel assemblies (1D tunnels, layered structures such as 103 birnessite and buserite and 3D spinel tunnel structures) (72). Chen et al. (55) studied the 104 influence of tunnel structures of various manganese oxide catalysts (pyrolusite, cryptomelane and 105 todorokite) on HCHO oxidation. Cryptomelane displayed the highest activity with 100% HCHO 106 conversion at 140°C, 400 ppm HCHO concentration, and a space velocity of 18,000 mL/g·h, while 107 20 and 40% conversions were attained by pyrolusite and todorokite, respectively, under similar 108 109 reaction conditions. Tunnel size and structure were shown to be the determinant factors affecting 110 activity other than factors such as specific surface area, degree of crystallinity, surface reducibility and average oxidation states of the catalysts. Similarly, Zhang et al. (19) attributed the high HCHO 111 catalytic activity of δ -MnO₂ to its interlayer and tunnel structures which help in expediting 112 113 adsorption and diffusion to and from the catalyst's active sites. These results indicate that activity is closely related to morphological and structural properties of the catalyst. 114

115 Tian et al. (54) investigated the impact of synthesis temperature on birnessite type manganese oxide catalysts. A general trend of catalytic activity rise was observed with increasing synthesis 116 temperature from 80 to 120°C. The catalyst synthesized at 120°C exhibited the highest activity 117 118 with a HCHO conversion of 100% at 100°C. This observed high activity was attributed to higher surface reducibility, specific surface area, crystallinity and porosity compared to other catalysts 119 synthesized at different temperatures. Wang et al. (73) studied the effect of H₂O molecules on the 120 activity of birnessite manganese for HCHO oxidation at room temperature and indicated that 121 activity depends on the amount of both adsorbed H₂O molecules and interlayer hydroxyl and H₂O 122 molecules present in the catalyst. Interestingly, as the drying temperature increases from 30 to 123 500°C, so does the specific surface area, but activity dramatically decrease as displayed in Figure 124 1(a) due to the reduction in water content. This result indicates that the activity of birnessite is 125

more a function of the interlayer H₂O molecule content than specific surface area. The presence of
 H₂O molecules enhances adsorption of HCHO molecules on the surface, conversion and
 desorption of intermediates from the catalyst's surface.

Furthermore, the effect of manganese vacancy (V_{Mn}) on the activity of birnessite-type MnO₂ was 129 studied (74). Their results indicate that the presence of V_{Mn} improves the content of the surface 130 adsorbed oxygen containing species facilitated by the presence of interlayer K⁺, which helps in 131 charge imbalance compensation caused by the created vacancy. The effect of surface pores created 132 by the modification of birnessite-MnO₂ with nitric acid and tetra-ammonium hydroxide on HCHO 133 oxidation was also investigated (75). The created defects as proposed by the authors served as sites 134 for activation of molecular oxygen and H₂O, hence resulting in the increased density of active 135 136 surface oxygen species and activity of the modified catalyst.

Tian et al. (56) also examined the effect of synthesis reaction temperature on the catalytic activities 137 of various cryptomelane manganese octahedral molecular sieve K-OMS-2 catalysts. The K-OMS-138 2 nanoparticle catalyst prepared at room temperature exhibited a higher catalytic activity compared 139 to K-OMS-2 nanorod structured catalyst prepared at 100°C. The former attained 64% HCHO 140 conversion at the reaction temperature of 100°C was due to the existence of a higher proportion of 141 pore channels compared to the 10% accomplished by the latter under the same reaction conditions. 142 Tian et al. (36) studied the relationship between textural properties and catalytic activity of the 143 cryptomelane manganese oxide catalyst, and showed that a resultant increase in textural properties 144 (specific surface area and pore volume) as a result of synthesis temperature increment from 15 to 145 70°C led to a dramatic improvement in catalytic activity. 146

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Three dimensional mesoporous MnO₂ (3D MnO₂) catalysts prepared via nanocasting route using 147 KIT-6 hard template, which retain the mesoporous properties of the template were shown to be 148 promising for HCHO oxidation (76). The mesoporous structure, high specific surface area and 149 large number of surface active Mn⁴⁺ ions enhanced its high activity compared to the corresponding 150 α -MnO₂ and β -MnO₂ nanorod catalysts. Complete HCHO oxidation to H₂O and CO₂ was 151 152 accomplished at 130°C with 3D MnO₂ while same conversion was achieved at 140°C and 180°C respectively with α -MnO₂ and β -MnO₂ under similar reaction conditions. This indicates that 153 activity is closely related to morphological structures of the catalyst. The catalytic activities of α , 154 155 β , γ and δ phase structures of MnO₂ were also investigated (19). The following order of activity for complete HCHO oxidation was established: $\delta > \alpha > \gamma > \beta$ - MnO₂. Complete HCHO oxidation 156 was attained at 80°C on δ -MnO₂. Its high activity was suggested to be a function of its abundant 157 concentration of lattice oxygen species (see section 4.3 for further discussion on the roles active 158 surface oxygen species) and 2D layered structure, the latter of which enables easy adsorption and 159 diffusion of HCHO. 160

Chen et al. (21) prepared mesoporous hollow and honeycomb structured K_xMnO₂ nanospheres. 161 The former displayed higher activity of up to 100% HCHO removal at 80°C, while the latter 162 attained similar removal efficiency at 100°C under similar reaction conditions. Catalytic activity 163 of hollow structured K_xMnO₂ nanospheres was attributed to its porosity and the ability to retain 164 HCHO in its pores for a longer period compared to honeycomb structured catalyst. Zhou et al. (77) 165 166 examined the catalytic activity of various structures of Mn and the following order of decreasing activity was established at reaction temperature below 120°C: cryptomelane Mn₂O > birnessite 167 $Mn_2O > ramsdellite Mn_2O > monoclinic MnOOH$. However, birnessite turns out to exhibit better 168 activity for complete HCHO oxidation at higher temperature, achieving 100% removal at 140°C 169

while cryptomelane at 160°C under similar conditions. This was possibly a result of weakened
HCHO adsorption on cryptomelane at lower temperature as suggested by the authors.

172 2.1.2 Cobalt oxide based catalysts

173 Cobalt oxide is another important transition metal-based oxide that has recently been attracting attention for low temperature VOC catalytic applications. Similar to MnO_x, Co₃O₄ also exhibit 174 175 various morphologies which include nano-fibers (78), nano-sheets, nano-cubes, nano-rods (79) 176 and the morphologies influence its catalytic activity through the exposure of catalytically active surface sites (70). The activity of Nano-, 2D- and 3D-Co₃O₄ structures for complete HCHO 177 178 oxidation was compared and the following order of reactivity was established: $3D-Co_3O_4 > 2D Co_3O_4 > nano-Co_3O_4$ (70). The superior activities of 3D and 2D-Co_3O_4 was ascribed to their 179 mesoporous channel structure which enables easy diffusion of reactants to undergo reaction on the 180 active surface compared to the non-porous nano-Co₃O₄; in addition, the best performance of 3D-181 Co₃O₄ was closely related to its abundant surface adsorbed oxygen species, large specific surface 182 area and exposed active Co^{3+} species on the (2 2 0) crystal face. Similarly, Ma et al. (51) indicated 183 that 2D-Co₃O₄ composed mainly of active Co^{3+} species on the (1 1 0) facet was able to achieve 184 20.3% HCHO oxidation at room temperature. 185

Fan et al. (67) investigated the effect of precipitants on the activity of Co₃O₄ catalysts. The catalysts
synthesized using carbonates and bicarbonates displayed better textural properties and superior
catalytic activities. The catalyst produced from KHCO₃ reached 100% HCHO conversion at 90°C,
compared to 120°C and 130°C for those prepared from KOH and NH₃·H₂O, respectively.
Recently, Wu et al. (78) described the utilization of porous Co₃O₄ nanofibers prepared by spiral
electrospinning and controlled calcination (500°C) as highly active catalysts for HCHO oxidation.
Complete oxidation was accomplished at 98°C under a space velocity of 30,000 mL/g·h and it was

shown to be highly stable for up to 160 hrs. Its activity was attributed to its high specific surfacearea and large pore volume which provided more active sites for the reaction.

Despite the reported influence of morphological structure on the activity of Co₃O₄ based catalysts, very few structures have been exploited for HCHO oxidation. There is therefore the need for further work to investigate other structures such as nano-sheets, nano-cubes, nano-rods and the extent to which they can influence other properties such as specific surface area, porosity and exposure of active metal sites and active oxygen species for improved low temperature HCHO oxidation.

201 2.1.3 Other metal oxides

Beside manganese and cobalt based catalysts, other metal oxides have been investigated in the 202 203 literature. Xia et al. (69) synthesized 3D ordered rhombohedra Cr₂O₃ using assisted ultrasound nanocasting in the presence of 3D mesoporous silica (KIT-6) as template. The catalyst attained up 204 to 90% HCHO conversion at 117°C. The ultrasound assisted synthesis played a major role in 205 improving the characteristics and activity of the catalyst compared to similar catalysts synthesized 206 in the absence of ultrasound. Huang et al. (80) first recounted the use of bifunctional Eu-doped 207 CeO₂ with both thermal- and photo-catalytic oxidation capabilities for HCHO oxidation. Doping 208 Eu onto CeO₂ dramatically enhanced its HCHO oxidation activity and complete oxidation 209 temperature was reduced from 310°C on pure CeO₂ to 120°C on 4% Eu doped CeO₂. Catalytic 210 activity was promoted through oxygen vacancy creation on the surface of CeO₂, greater redox 211 ability, more abundant surface active Ce³⁺ and enhanced surface reaction. The created defects 212 provide more sites for oxygen activation hence providing more active surface oxygen species for 213 214 reaction. In addition, the catalysts exhibited an outstanding stability of up to 100 hrs on stream without any sign of deactivation. Similarly, Zeng et al. (81) demonstrated that hydrogenation 215

treatment of TiO₂ and C-TiO₂ improved their activity through the creation of more surface oxygen vacancy and surface hydroxyl groups. The untreated catalysts virtually displayed no activity for HCHO oxidation in a static chamber reaction after 4 hrs at room temperature in the absence of light illumination, while the treated catalysts including H-TiO₂ and H-C-TiO₂ were able to respectively attain 53% and 57% removal under similar reaction conditions.

In general, manganese and cobalt based catalysts were demonstrated to be active compared to other transition metals catalysts for HCHO oxidation. However, not much work has been done on other transition metals. In order to explore their full potentiality as viable catalysts, further work is needed to explore their utilization for low temperature HCHO oxidation. In addition, more work is required to explore surface oxygen vacancy or defects creation using various dopants to improve oxygen activation ability of metal oxide catalysts, to enrich active surface oxygen concentration for enhanced low temperature reaction.

228 2.2 Transition metal-based composites/mixed oxides

229 Transition metal-based composites have been widely applied in the control of pollutants such as CO (83,84) and ammonia (85). Composite catalysts for HCHO oxidation have also been 230 investigated and reported in the literature. These catalysts are generated by co-synthesis of two 231 transition metals oxides or the addition of other metal oxides either by co-precipitation (48,86) or 232 by other synthesis methods such as nanocasting (51). Composite catalysts were shown to exhibit 233 superior catalytic activities compared to the corresponding single materials synthesized using 234 similar procedure. This is due to synergistic or promotional influence of improved oxidation 235 capabilities either through higher surface oxygen mobility, creation of more oxygen vacancies 236 237 (48,86) or enhancing charge transport during redox cycles (87). To improve oxygen vacancy formation in a composite catalysts and reduce the energy requirement for such, it was suggested 238

that the dopant should possess weaker M-O bonds; have larger radius and possess lower
electronegativity (88). Various composites such as MnO_x-CeO₂ (48,86), Co₃O₄-CeO₂ (51), MnO₂-

241 Fe₂O₃ (89), CuO-MnO₂ (51,89), MnO_x-SnO₂ (90), Co₃O₄-ZrO₂ (91) and Co-Mn oxide (68) have

- 242 been studied for HCHO oxidation.
- 243 $2.2.1 MnO_x$ -CeO₂ composites

244 The composites of Mn and Ce oxides were shown to be active for HCHO oxidation owing to the synergistic effect of Mn high activity and O₂ storage capacity of ceria. In addition, the composites 245 aid in attaining higher oxidation states for Mn (83,92) which is vital for HCHO oxidation (87). 246 Formation of solid solution between Mn and Ce is also critical to achieving synergistic influence 247 through O₂ transfer mechanism (48,86). Solid solution was shown to be attained in the Mn-Ce 248 molar ratio (Mn/(Mn + Ce)) range of 0.3 to 0.5, with 0.5 molar ratio being the optimum 249 250 composition (48,86,93) and the solubility limit for the substitution of Ce ions by Mn ions in the composite (53). 251

252 Tang et al. (86) indicated that the manifestation of synergy in MnO_x-CeO₂ solid solution composite, which enables the composite to attain complete HCHO conversion at lower 253 temperature (100°C) compared to pure MnO_x and CeO₂. The synergy was attained through a series 254 of redox cycles (Mn^{4+}/Mn^{3+}) and Ce^{4+}/Ce^{3+} involving the activation of molecular oxygen by Ce 255 and its transfer to Mn. The optimum calcination temperature for solid solution stability was shown 256 to be 550°C, above which activity decreased owing to phase segregation. In another investigation, 257 Tang et al. (48) further highlighted that above Mn-Ce molar ratio of 0.5, MnOx crystallizes out of 258 the solid solution, with a consequent drastic reduction in catalytic activity, which is in conformity 259 260 with the findings of Li et al. (93). However, higher total decomposition temperature of 270°C was reported by Li et al. (93) for Mn_{0.5}Ce_{0.5}O₂. Nonetheless, the catalysts were able to totally oxidize 261

HCHO at room temperature in the presence of 506 ppm ozone (O₃) in the feed stream. The ozone molecules enhance the reaction by dissociating on the catalyst's surface and providing enough atomic oxygen species to drive the reaction (93).

265 Quiroz et al. (53) studied the impact of acid treatment over MnO_x-CeO₂ composite for HCHO oxidation. Their results indicated that the treatment had no pronounced effect on the textural and 266 267 redox properties of the composites within the solubility limit of Mn (Mn molar ratio ≤ 0.5) in CeO₂. However, at higher Mn molar ratio (0.7 and 1) when MnO₂ began to crystallize out of the 268 solid solution, significant improvement in catalytic activity was observed owing to the increased 269 270 specific surface area and higher oxidation state of surface Mn. It is important to mention that while textural properties of the solid solution were not altered by the acid treatment, the oxygen transfer 271 ability of CeO₂ was significantly hampered by the formation of Ce(SO₄)₂ resulting into a drastic 272 reduction in activity. Therefore, acid treatment has positive effects on pure MnO₂ and an inhibitive 273 effect on the synergy of the solid solution composites. The modification of birnessite-structured 274 MnO₂ with ceria for HCHO oxidation was also reported in the literature (94). The incorporation 275 of cerium ions inhibited the growth of MnO₂ crystals leading to particle size reduction and increase 276 in specific surface area, as the amount of doped ceria increases. The doping led to the increase in 277 278 the amount of oxygen vacancies and the rate of molecular oxygen activation into surface adsorbed oxygen species (O_2^- , O^- or terminal hydroxyl (OH) group (74)). The catalyst with the highest ratio 279 of surface adsorbed oxygen to lattice oxygen exhibited the best catalytic performance. At higher 280 281 ceria doping (Ce-MnO₂, molar ratio 5:10), the birnessite structure collapsed leading to decrease in catalytic activity. 282

Tang et al. (48) further demonstrated that impregnating Pt on MnO_x-CeO₂ composite significantly
improved its activity by attaining 100% HCHO conversion at room temperature as shown in Table

1. It is noteworthy to mention that the formed composite (Pt/MnO_x-CeO₂) was prone to 285 deactivation with increasing HCHO concentration. Its activity dropped from 100% conversion at 286 30 ppm to 54% in the presence of 580 ppm HCHO at room temperature as shown in Figure 2. 287 Meanwhile the same authors (86) showed that MnO_x-CeO₂ composite was stable for complete 288 HCHO oxidation over 48 hrs without any sign of deactivation at similar concentration (580 ppm). 289 It should however be noted that the latter experiment was conducted at 100°C and the high 290 temperature could aid the decomposition of intermediate species while the former was conducted 291 at room temperature. This indicates that Pt might be prone to deactivation by HCHO at higher 292 293 concentration, which needs to be investigated further.

294 *2.2.2 Co₃O₄-CeO₂ composites*

Ma et al. (51) investigated the oxidation of HCHO at room temperature on 2D-Co₃O₄-CeO₂ 295 composites. They discovered that no synergy or activity improvement was realized by the formed 296 2D-Co₃O₄-CeO₂ composites and that 2D-Co₃O₄ was more active than the composites with 297 respective HCHO conversions of 13.2% and 20.3% at room temperature. It should however be 298 noted that no solid solution formation was observed in the 2D-Co₃O₄-CeO₂ composites, which is 299 a phenomenon reportedly required for achieving synergy through ceria O₂ transfer mechanism 300 301 (48,86). The incorporation of Au in the composite (2D-Au/Co₃O₄-CeO₂) improved the composite's activity to 50% conversion at room temperature (Table 1) by promoting desorption of surface 302 active oxygen species (51). Liu et al. (50) presented the evidence of solid solution formation in 3D 303 304 ordered macroporous (3DOM) CeO₂-Co₃O₄ catalysts with low loading of Co₃O₄. However, the 3D-CeO₂-Co₃O₄ catalysts with various Ce-Co molar ratios were shown to possess low catalytic 305 activities. Loading Au particles (3 wt. %) on the composites distinctively improved their catalytic 306 activities, and in fact all the Au supported composites were able to completely convert HCHO into 307

308 CO₂ and H₂O below 65°C. The composites (3D Au/CeO₂-Co₃O₄), with lower Co content in CeO₂-309 Co₃O₄ in the region solid solution, presented better activity in this study. When higher Co molar 310 ratios was employed, segregation between CeO₂ and Co₃O₄ occurred, which led to weaker 311 interaction and thus reduced the activity of the Au/CeO₂-Co₃O₄ catalysts.

312 2.2.3 Co-Mn oxides Composites

The composites of Co and Mn oxides were also studied and shown to be more active than the 313 individual pure oxides for HCHO complete oxidation (95). Favorable synergy was obtained due 314 to the abundant surface adsorbed oxygen species generated by the oxygen vacancies created in the 315 Co-Mn oxides solid solution. Textural and redox properties were also shown to be affected by the 316 synthesis method employed. Co-precipitation synthesis produced a more active catalyst with 317 318 higher specific surface area, higher relative content of surface adsorbed oxygen species and surface active manganese (Mn⁴⁺) that completely oxidized HCHO at 75°C compared to 100°C over that 319 synthesized using sol-gel method. 320

Wang et al. (68) further studied the efficiency of the cycling "storage-oxidation" process for 321 HCHO oxidation using 3D ordered mesoporous Co-Mn oxides composite (3D-Co-Mn) 322 323 synthesized using KIT-6 template and that of Co-Mn oxides composite synthesized using coprecipitation. During the storage process, the catalysts were first exposed to a feed stream of O₂ 324 and HCHO; subsequently, the stored HCHO was oxidized in a flow of O₂ at elevated temperature 325 to evaluate its activity. The 3D-Co-Mn oxides composite was shown to have superior storage 326 capacity of up to 0.8 mmol/g-catalyst and HCHO complete oxidation temperature of 70°C, 327 compared to the storage capacity of 0.428 mmol/g-catalyst and HCHO complete oxidation 328 temperature of 75°C respectively achieved by co-precipitated Co-Mn oxides composite. The 329

excellent activity of 3D-Co-Mn oxides composite was attributed to its larger specific surface area and ordered mesoporous structure. Lu et al. (96) studied HCHO oxidation over MnO_x -Co₃O₄-CeO₂ composites and found out that the incorporation of MnO_x into Co₃O₄-CeO₂ greatly improved its textural properties and the amount of available surface active oxygen species, resulting in an improvement in catalytic activity.

335 2.2.3 Other Composites

336 Other composites such as MnO_x-SnO₂ (90) and zirconia supported cobalt oxide catalysts (91) have been reported in the literature. Wen et al. (90) showed that the redox properties of MnO_x-SnO₂, 337 338 which were in turn influenced by the preparation method, played a crucial role in determining its activity. Their experimental results highlighted that higher oxidation states of Mn were more active 339 in HCHO oxidation. The dominant oxidation state of Mn in the composite synthesized using redox 340 co-precipitation was Mn⁴⁺ while Mn³⁺ was the main state in the co-precipitated composite, which 341 accounted for the difference in their activities as shown in Table 3. This is in agreement with other 342 reported literatures for the active state of Mn for HCHO oxidation (76,86). Lu et al. (87) 343 investigated the catalytic performance of graphene-MnO₂ (G-Mn) hybrid for complete oxidation 344 of HCHO. The G-Mn hybrid catalyst achieved complete HCHO conversion at 65°C, compared to 345 140°C for pure MnO₂ and the completely inactive graphene (G) nanosheets. The hybrid system 346 exposed more Mn⁴⁺ active sites, enhanced charge transport during Mn redox cycle and offered a 347 larger amount of surface HYDROXYL species which eased HCHO oxidation and improved 348 349 activity.

350 **3. Catalyst immobilization on porous materials for practical HCHO oxidation**

For practical applications, deployment of powdered form materials especially nanosized, gives rise
to engineering challenges such as dust contamination and nanoparticle leaching in flue gas streams

(82). Therefore, catalysts immobilization on porous materials with low air pressure drop is 353 considered viable for air purification processes. A variety of porous materials with low air 354 resistance including polyethylene terephthalate (PET) (18), polyester particulate filter (97) and 355 porous cellulose fiber (82) have been used as supporting materials for immobilizing nanoparticle 356 catalysts for HCHO oxidation. A summary of the activity of these immobilized catalysts is 357 358 presented in Table 4. Wang et al. (18) reported the in-situ coating of δ -MnOx nanosheet on the surface of PET first through surface reaction followed by in-situ deposition. The formed composite 359 which is light with low air resistance and high specific surface area, proved highly active and stable 360 for low concentration HCHO (0.6 mg/m^3) oxidation at room temperature for 10 hrs. Sidheswaran 361 et al. (97) supported Mn based catalyst with 84% nsutite, 2% cryptomelane and 13% manjiroite 362 composition on the surface of a thin polyester particulate filter for heating ventilation and air 363 conditioning (HVAC) and evaluated its performance for HCHO oxidation. Experimental results 364 indicated that stable single pass HCHO oxidation with over 80% removal efficiency at room 365 temperature was achieved continuously for 35 days for both high and low face velocities close to 366 typical building air ventilation systems. 367

Zhou et al. (82) likewise demonstrated an in situ deposition of MnO₂ nanosheets on cellulose fiber 368 369 composite (8.86 wt.% MnO₂/cellulose fiber) and showed that even though birnessite powder is slightly more active than the composite (100% and 99.1% at 140°C, respectively), the composite 370 is about 19 times more active in terms of HCHO removal per mg of MnO₂. However, the observed 371 372 difference could supposedly be attributed to the removal/adsorption capacity of cellulose fiber and not necessarily HCHO conversion as CO₂ generation and catalytic activity of the cellulose fiber 373 were not monitored in the experiment. Li et al. (98) and Dai et al. (71) respectively utilized AC for 374 immobilizing birnessite-MnO₂. In both cases, HCHO was completely converted into CO₂ in a 375

static reaction chamber. However, the conversion patterns in the individual experiments were such 376 that the concentration of HCHO sharply dropped in the first 60 mins with little corresponding CO₂ 377 generation, possibly indicating that the molecules were actually adsorbed onto the surface of AC. 378 The adsorbed molecules were consequently and slowly converted into CO_2 over 6 hrs (98) and 9 379 hrs on birnessite-MnO₂/AC (71) respectively as shown in Figure 3(a&b). Similar patterns were 380 381 observed over unsupported birnessite as shown in Figure 1(b) (73). In contrast, simultaneous HCHO conversion and CO₂ generation were observed over modified birnessite with manganese 382 vacancies and up to 81.7% conversion was attained in the first one hour of the static experiment 383 384 (74). This indicates that the HCHO is instantaneously converted into CO₂, as opposed to the AC supported birnessite (71,98) in which the HCHO is adsorbed onto the AC and subsequently 385 converted over time. Few immobilized transition metal-based catalysts for HCHO oxidation have 386 been reported as presented in Table 4; therefore, further investigations are required to evaluate 387 more effective substrate materials for immobilization and to understand their interactions with the 388 catalysts and effectiveness for practical application in air purification process. In addition, these 389 materials need to be tested under indoor conditions in air purifiers or HVAC systems and evaluate 390 the effect of conditions such as particle leaching, dust contamination, relative humidity and 391 392 temperature variation, on the effectiveness and stability of these materials.

4. Influence of catalyst based factors on HCHO oxidation efficiency

The effectiveness of HCHO oxidation process is majorly related to the activity/reactivity of the catalyst deployed and other reaction parameters such as temperature, concentration, space velocity, catalyst mass and relative humidity. Catalytic activity is in turn influenced by a number of physicochemical properties which include structure and morphology, preparation or synthesis method, degree of crystallinity, surface reducibility, specific surface area, amount of active oxygen species and active metal sites (54). A combination of these properties is decisive for high catalytic
activity and for an effective HCHO oxidation at low reaction temperature.

401 4.1 Influence of synthesis methods and conditions on textural properties, morphology and

402 activity

Several conventional preparation methods can be used to synthesize catalysts for HCHO oxidation 403 404 including sol-gel method (56), precipitation and co-precipitation (86,97). Recent researches have mainly focused on improving and modifying catalysts preparation methods to enhance their 405 performance and catalytic activities through the utilization of synthesis techniques such as 406 407 hydrothermal synthesis (77), electrospinning (78), electrodeposition (80) and hard template nanocasting (70,76). Synthesis methods and reaction conditions are able to tailor catalysts' textural 408 409 and surface properties, structures and morphologies (as shown in Figure 4 (77)) and hence influence their catalytic activity (36,54) as shown in. For instance, the utilization of ultrasound 410 assisted nanocasting of 3D-Cr₂O₃ using KIT-6 as template helps in enhancing the penetration of 411 precursor materials into the mesoporous structure of KIT-6, which proved efficient in improving 412 its specific surface area, pore volume and mesoporousity compared to 3D-Cr₂O₃ synthesized 413 without the aid of ultrasound (69). Zhang et al. (19) highlighted that by varying reaction conditions 414 415 in a hydrothermal synthesis, various crystal structures of MnO₂ including: α - and δ -MnO₂; β - and γ -MnO₂, with entirely different properties can respectively be produced from the same starting 416 materials. 417

Reaction temperatures in hydrothermal synthesis have tremendous effects on morphology and structural evolution of structured manganese catalysts. Tian et al. (54) highlighted that the reaction temperature greatly affects the crystallinity, surface reducibility, specific surface area and activity of birnessite structured manganese oxide catalysts. At low synthesis temperature of 80°C, poorly

crystalline birnessites were produced, whilst the birnessites were transformed into well-crystalline 422 structures when reaction temperature was increased to 100°C, thereby increasing its activity. 423 However, higher temperature (140°C) led to decomposition of the birnessite structure as a result 424 of excessive reduction of Mn⁷⁺ by benzyl alcohol during the synthesis. Zhou et al. (77) also 425 observed that birnessites structure obtained at 120°C synthesis temperature transformed into 426 427 monoclinic MnOOH at 150°C and eventually collapsed into 1D MnOOH nanorods at 180°C as shown in Figure 4. Similarly, a correlation between synthesis temperature and catalyst's 428 morphology for cryptomelane-type manganese octahedral molecular sieve (K-OMS-2) was 429 430 reported by Tian et al. (56). The morphologies of the K-OMS-2 catalysts changed from nanoparticles at room temperature to nanorods at reaction temperatures of 80-100°C and to 431 nanowires at 120°C. 432

Tian et al. (36) reported that essential properties of cyptomelane catalyst such as morphology, 433 crystallinity, specific surface area and pore structure could be tailored by manipulating the 434 hydrothermal synthesis temperature. Increasing synthesis temperature distinctively raised the 435 crystallinity and textural properties of the catalyst. However, too high temperatures diminished 436 crystallinity and this is in agreement with the findings of Tian et al. (54) for birnessite manganese 437 438 catalysts. Textural properties such as specific surface area and pore volume were also shown to vary with the synthesis temperature. An increase in the synthesis temperature led to a dramatic 439 improvement in textural properties: specific surface area and pore volume increased from $68 \text{ m}^2/\text{g}$ 440 and 0.2 cm³/g at synthesis temperature of 15°C, to 206 m²/g and 0.3 cm³/g at 70°C, respectively. 441 Such an improvement in textural properties in turn led to enhancement in catalytic activity. 442

Calcination temperature was also demonstrated to be very critical for catalysts structural stability
and activity. Wang et al. (73) indicated that between 30 and 300°C only a little change was

observed in the sizes of birnessite nanospheres and that the structure was prevented from 445 collapsing by interlayer K⁺ and H₂O molecules present. However, when the calcination 446 temperature reached up to 500°C, the layered structure of birnessite completely collapsed and 447 transformed into cryptomelane with a 2 \times 2 tunnel structure due to the loss of interlayer H₂O 448 molecules. The calcination temperature of composite MnOx-CeO₂ prepared using modified co-449 precipitation was shown to greatly influence its catalytic activity for HCHO oxidation (86). As the 450 temperature rises from 300-500°C, the relative amount of surface active Mn⁴⁺ and lattice oxygen 451 also rises, however at higher temperature (700°C) both the activity and the relative amount of 452 453 lattice oxygen dramatically decreased due to phase segregation evident by the appearance of MnO₂ crystals in the otherwise solid solution composite. Likewise, Wu et al. (78) indicated that high 454 calcination temperatures (600-800°C) led to the decrease in the activity of Co₃O₄ nanofibers as a 455 result of aggregation growth and collapse of the catalyst's mesoporous structure. 456

The utilization of templates and surfactants in catalysts synthesis for HCHO oxidation is also 457 attracting attention. When templates are deployed in catalyst preparation, the catalyst's structures 458 and morphology could easily be manipulated to enhance their catalytic activities (56,76) and 459 replicate the characteristics of the template materials used. While Shi et al. (95) demonstrated that 460 461 Co-Mn oxides composites prepared using co-precipitation method were far more active than those prepared using sol-gel method, the same group (68) further disclosed that 3D mesoporous Co-Mn 462 oxides synthesized using KIT-6 hard template possessed superior activity owing to their better 463 textural properties. Similarly, the synthesis of 3D-Cr₂O₃ (69), 3D-Co₃O₄ (70) and 3D-MnO₂ (76) 464 have been demonstrated using KIT-6 molecular sieve hard template for HCHO oxidation. The 465 presence of the template improved the mesoporous structure and specific surface area and exposes 466 more active sites thereby enhancing catalytic activity. Other mesoporous structures including 2D-467

MnO₂ (99) and 2D-Co₃O₄ (51,70) have been prepared using SBA-15 hard templates, which turned out to be more active than the corresponding non-porous materials but less active than their 3Dcounterparts. Furthermore, Tian et al. (56) highlighted that highly active nanoparticle cryptomelane (K-OMS-2) catalyst could be synthesized at room temperature with the aid of cetyltrimethylammonium bromide (CTAB) surfactants without necessarily the use of high reaction temperature as earlier reported (*36*).

474 Acidity of the reaction medium and the type of acid deployed in controlling the pH have significant effects on the morphology of manganese based catalysts. Chen et al. (21) showed that mesoporous 475 476 hollow and honeycomb K_xMnO₂ nanospheres with varying properties and activities could be produced by varying the ratio of KMnO₄ and oleic acid during the synthesis process. Zhou et al. 477 (77) revealed that introducing H₂SO₄ (decreasing the pH) into a reaction medium that would 478 otherwise produce birnessite structured manganese led to the production of cryptomelane 479 structured manganese. However, the type of anions present in the acid is also critical in 480 determining the resultant morphology under similar conditions. While the addition of H₂SO₄ led 481 to cryptomelane structure, addition of H₃PO₄ and HNO₃ produced birnessite and ramsdelite 482 structured manganese, respectively as shown in Figure 4. This shows that different anions play 483 dissimilar roles in channeling the mechanism of crystal formation towards a particular 484 morphology. 485

486 4.2 Influence of catalyst's morphology on activity

487 Catalysts' morphology and structure play an important role in promoting catalytic activities.
488 Control over the morphological structure of catalysts is able to influence other properties such as
489 pore size, specific surface area and the exposure of available surface active sites. The activity of
490 manganese oxide catalysts with square tunnel structures (pyrolusite, cryptomelane and todorokite)

is majorly affected by their tunnel sizes (*55*). Similarly the available surface active sites depend on the nature of the catalysts. 3D- MnO₂ catalyst have more exposed Mn⁴⁺ on its crystal lattice plane thereby improving its activity compared to the one-dimensional α -MnO₂ and β -MnO₂ nanorod materials (*76*).

Chen et al. (55) investigated the effect of MnO_x tunnel structure on the active oxidation of HCHO 495 as shown in Figure 5. Three types of Mn tunnel structures were investigated including pyrolusite 496 with tunnel structure of ca. 0.23×0.23 nm², cryptomelane consisting of double edge sharing MnO₆ 497 octahedral with tunnel diameter of ca. 0.46×0.46 nm² and todorokite composed of triple chains 498 edge-sharing MnO₆ with tunnel of ca. 0.69×0.69 nm². Experimental results indicated that 499 cryptomelane had the highest catalytic activity achieving 100% HCHO conversion at 140°C. 500 501 Cryptomelane possesses an effective tunnel diameter close to HCHO's dynamic diameter (0.234 nm) thereby allowing better adsorption and higher catalytic activity (55). Yu et al. (41) indicated 502 503 that the micropore channel structure of nest- and urchin-like MnO₂ improved their catalytic activity 504 by allowing better adsorption of HCHO compared to that of cocoon-like MnO₂ without mesopores 505 present despite having higher specific surface area.

Tian et al. (54) also highlighted that the presence of smaller catalytic pore sizes could extend the 506 residence of HCHO molecules in the catalyst's pores thereby improve its oxidation efficiency. 507 508 Correspondingly, the work of Tian et al. (56) underscored that the pore structures of nanoparticle K-OMS-2 were readily accessible to HCHO molecules, enhancing higher adsorption and better 509 activity compared to K-OMS-2 nanorods with similar specific surface areas. Zhang et al. (19) 510 further highlighted that the 2D layered tunnel structure of δ -MnO₂ enhanced its activity through 511 512 facilitating the adsorption and desorption of HCHO to active sites compared to other structures (α -, β - and γ - MnO₂). In another research, Chen et al. (21) stressed that catalyst's activity depends 513

more on its porosity and nanoplatelets crystal size than specific surface area. Mesoporous hollow
K_xMnO₂ nanospheres, which have smaller nanoplatelets size and specific surface area twice less
than those of its corresponding mesoporous honeycomb structure, achieved higher activity in
HCHO oxidation.

Wang et al. (73) showed that the structure of birnessite is such that it contains HYDROXYL groups 518 and H₂O molecules at the interlayer surface. The presence of these molecules greatly improved its 519 catalytic activity for HCHO removal even at room temperature. After the removal of both the 520 hydroxyl groups and H₂O molecules at high temperature, the birnessite structure collapsed and the 521 522 catalyst lost its high activity. Likewise, Sidheswaran et al. (97) indicated that the existence of H₂O molecules in interstitial voids of MnO₂ based catalysts with varying composition of nsutite, 523 cryptomelane and pyrolusite contributed to its higher catalytic activity compared to 99% pyrolusite 524 catalyst which has smaller tunnel structure and lower H₂O molecules content in its interstitial 525 voids. 526

527 4.3 Roles of active surface oxygen in HCHO oxidation

It is reported in the literature that surface oxygen species either in the form of surface adsorbed 528 oxygen species (O_2^- , O^- or terminal hydroxyl (OH) group (74)) and lattice oxygen are critical to 529 530 the effectiveness of the catalytic oxidation of HCHO and other intermediates into CO₂ and H₂O (51,67,95). However, there is a discrepancy regarding the respective roles of each type of oxygen 531 specie for HCHO oxidation over transition metal based catalysts. For example, Zhang et al. (19) 532 stipulated that since HCHO oxidation was shown to conform to Mars van Krevelen mechanism, 533 abundant lattice oxygen on the catalyst surface will lead to higher catalytic activity. They further 534 showed that the catalytic activities of four MnO₂ catalysts (α , β , γ and δ -MnO₂) are closely linked 535 to their respective amount of surface lattice oxygen (relative to surface adsorbed oxygen species) 536

and the following order of activity was established: $\delta - > \alpha - > \beta - > \gamma - MnO_2$. Similarly, Tang et al. (86) indicated that the catalytic activity trend of MnO_x-CeO₂ composite catalysts synthesized using various methods, followed the trend of their relative lattice oxygen content. Those richer in lattice oxygen relative to surface adsorbed oxygen presented better activity.

On the other hand, surface adsorbed oxygen species were shown to be directly involved in the 541 542 activation of HCHO and its subsequent oxidation into CO₂ and H₂O (74,75). It was shown that the activity of layered birnesite-MnO₂ catalysts with surface pits corresponds to the relative amount 543 of surface adsorbed oxygen species on the catalysts (75). The surface pits acts as vacancies for 544 activation of molecular oxygen and or H₂O into surface adsorbed oxygen and the higher the 545 relative amount of these species the better the catalytic activity. Wang et al. (94) contended that 546 547 while surface adsorbed oxygen species participates in the oxidation reaction of HCHO, lattice oxygen species enhance their formation through its complex interaction with oxygen vacancy and 548 molecular oxygen. Huang et al. (80) reported that the enhancement of the relative amount of 549 surface adsorbed active oxygen species was achieved through the creation of surface defects by 550 doping Eu on CeO₂. The created oxygen vacancies served as sites for oxygen activation into active 551 superoxide (O²⁻) species which could partake in surface reaction, hence enhancing the 552 concentration of surface adsorbed oxygen and the catalyst's activity. Similarly, surface defects 553 created on birnesite-MnO₂ catalysts served as pits for activation of molecular oxygen and H₂O into 554 surface active oxygen species, hence increasing the surface density of these species leading to 555 556 higher catalytic activity (74,75). Ma et al. (51) also underscored the role of surface adsorbed oxygen (O²⁻, O⁻) in HCHO oxidation over cobalt based catalysts. They showed that catalytic 557 activity is closely associated to the ease with which active surface oxygen desorb from catalysts 558 surface. They further demonstrated that catalytic activity can be improved by increasing the ease 559

with which surface active oxygen desorb from catalyst's surface through the incorporation of Au nanoparticles into Co₃O₄-CeO₂ composite. Likewise, it was demonstrated that 3D-MnO₂ with abundant concentration of surface adsorbed oxygen species which easily desorb from the catalyst at lower temperature exhibited better catalytic activity compared to 2D- MnO₂ with lower surface adsorbed oxygen species concentration and nano-Co₃O₄ with virtually no active oxygen species (O^{2-}, O^{-}) present (70).

In addition, Lu et al. (87) indicated that surface adsorbed oxygen species (hydroxyl group) and not 566 lattice oxygen are responsible for the high activity of graphene-MnO₂ hybrid catalyst for HCHO 567 568 oxidation. They further pointed out that the abundant relative amount of surface adsorbed hydroxyl group on the catalyst did not only improve its catalytic activity but also simplified HCHO 569 conversion pathway by direct hydroxyl ion oxidation of formate species to CO₂ without the 570 formation of CO intermediates. It was also shown that surface adsorbed hydroxyl groups were 571 572 directly involved in the room temperature oxidation of HCHO on the surface of birnessite, which led to the formation of formates and carbonates (73). Similarly, Wang et al. (18) showed that the 573 high amount of surface adsorbed oxygen in the form of hydroxyl species were responsible for the 574 575 high HCHO oxidation activity of δ -MnO₂/PET at room temperature. Similarly, Fan et al. (67) underscored that hydroxyl species on the surface of Co₃O₄ are the key active surface oxygen 576 required for the formation of dioxymethylene (DOM) species and their subsequent conversion to 577 formate species in HCHO oxidation process. 578

579 4.4 Influence of reducibility and active metal sites on catalysts surface

The surface reducibility of catalysts plays an important role in determining their catalytic activity.
It indicates the amount of reactive species on the catalyst's surface and their onset reaction

temperatures (54). It could also indicate catalyst's oxygen mobility, which might cause more

oxygen adsorption and further excitement to active oxygen species which participate in the 583 oxidation reaction (19). Tian et al. (54) indicated a correlation between reduction temperature and 584 catalytic activity for birnessite manganese based catalysts. The established trend disclosed that the 585 lower the catalyst's reduction temperature (higher surface reducibility), the higher its HCHO 586 oxidative activity and vice-versa. Tang et al. (86) presented that owing to solid solution formation, 587 MnO_x-CeO₂ catalysts possessed higher surface reducibility and better catalytic activity compared 588 to pure MnO_x and CeO_2 . Furthermore, the catalysts' activity reduced as surface reducibility 589 decreased (high reduction temperature) with increasing calcination temperature because of phase 590 591 segregation. Xia et al. (69) also indicated that chromia catalysts with the highest reducibility possessed the best HCHO catalytic activity. Similar trends was also reported for HCHO oxidation 592 over Co₃O₄ catalysts (67). 593

The active sites present on catalyst's surface are critical and key for determining activity. The 594 higher the concentration or availability of exposed active metal sites on catalyst's surface, the 595 better its HCHO oxidative activity. Bai et al. (70) presented that Co^{3+} ions are the surface active 596 state and sites of 3D-Co₃O₄ for HCHO oxidation. The (220) crystal plane of the catalyst is 597 majorly composed of Co^{3+} against the Co^{2+} present on the (111) crystal plane of nano-Co₃O₄. 598 thereby making it more active and achieving complete HCHO oxidation at lower temperature. 599 Similarly, Ma et al. (51) demonstrated that the (110) facet of 2D-Co₃O₄ which is mainly composed 600 of Co³⁺ is the key active facet for effective HCHO oxidation. High activity for HCHO oxidation 601 by 3D-MnO2 was attributed to the presence of high content of Mn⁴⁺ active sites on the exposed 602 (110) crystal plane surface (76). These exposed ions provided sufficient sites upon which the actual 603 oxidation reaction takes place, thereby improving its activity. It was also shown that Mn⁴⁺ is the 604 active state and site for HCHO oxidation for MnO_x-CeO₂ based composite and the catalyst with 605

the highest amount of Mn⁴⁺ possessed the best activity (48,86,87). Zhang et al. (19) also 606 highlighted that amongst their investigated catalysts (α , β , γ and δ -MnO₂), those with the highest 607 content of Mn⁴⁺ displayed the best HCHO catalytic activity. Improving the electrochemical 608 properties of catalysts was proven to enhance their catalytic activities. Lu et al. (87) established 609 that the charge transport ability and interfacial electron transfer of MnO_2 during Mn^{4+}/Mn^{3+} redox 610 cycle can be enhanced by incorporating graphene into MnO₂ (G-Mn hybrid), which was revealed 611 to improve the electrical conductivity of MnO₂ and decrease electron transfer resistance. In 612 addition, it offered high specific surface area and 2D plane structure for increased exposure of 613 metal surface active sites (100). 614

4.5 Influence of specific surface area on HCHO catalytic activity

Catalysts' specific surface area is an important parameter affecting HCHO catalytic oxidative. 616 High specific surface area could enhance adsorption of HCHO molecules onto the catalyst's 617 surface and help in exposing more surface active sites, improving surface lattice defect and oxygen 618 vacancy, thereby leading to abundant surface active oxygen species (70,78), which is a key 619 requirement for oxidation reactions. A correlation between catalyst's surface and activity was 620 established by Wu et al. (78) for Co₃O₄ nanofibers. Activity increase was observed with increasing 621 specific surface area and pore volume as the calcination temperature increased from room 622 temperature to 500°C, above which a decline in activity was observed due to the destruction of the 623 mesopores and the consequent decrease in specific surface area at higher temperatures. Tian et al. 624 625 (36) also demonstrated that higher specific surface area greatly enhanced catalytic activity for HCHO oxidation. Nanoparticle cryptomelane catalyst, with specific surface area as high as 206 626 m^2/g and smaller uniform mesopores exhibited higher catalytic activity than the corresponding 627 nanorod cryptomelane catalyst with a specific surface area of 68 m^2/g . The high specific surface 628

area and small nanoparticle sizes improved cryptomelane activity by increasing the available surface active sites and more accessible pore channels for adsorption and desorption of HCHO molecules and reaction products respectively (*36*). Sekine (*2*) showed that fine MnO₂ catalysts with higher specific surface area (163 m²/g) achieved higher HCHO oxidation compared to MnO₂ catalysts with a specific surface area of 61 m²/g. Similarly, Tian et al. (54) also reported that birnessite structured manganese catalysts having higher specific surface area exhibited better catalytic activities compared to those with lower specific surface areas.

However, high specific surface area does not always translate into superior catalytic activities in 636 637 HCHO oxidation reactions. The activity of birnessite-type MnO₂ was shown to be independent on the catalysts' specific surface area trend (74) but on the relative content of surface adsorbed oxygen 638 species and the manganese oxidation state. The catalyst with the highest manganese vacancy had 639 the lowest specific surface area but exhibited the best catalytic activity. Shi et al. (95) also 640 demonstrated that the activity trend of Co-Mn oxides composite with varying Co/Mn molar ratios 641 did not follow their specific surface area trend, and in fact the catalysts with the smallest specific 642 surface area exhibited the best catalytic activity. Complete oxidation of HCHO was achieved at 643 75°C on Co-Mn oxides (molar ratio: 3/1) with specific surface area of 92 m²/g while only 64% 644 conversion was attained on Co-Mn oxides (molar ratio: 2/1) with specific surface area of $172 \text{ m}^2/\text{g}$ 645 at the same reaction temperature. The higher activity of Co-Mn (3/1) oxides was a result of increase 646 in the relative amount of surface adsorbed oxygen species realized from the creation of solid 647 solution by the incorporation of Mn into the lattice structure of Co₃O₄. Fan et al. (67) also showed 648 that Co₃O₄ catalysts with high surface content of K⁺, CO₃²⁻ and Co³⁺ exhibited better activity 649 despite having relatively smaller specific surface area. 650

5. Effect of reaction conditions on HCHO oxidation efficiency

Reaction parameters for HCHO oxidation process such as temperature, relative humidity, reactants' space velocity and HCHO concentration play great roles in ensuring the overall effectiveness of the oxidation reaction. These parameters need to be optimized especially reaction temperature considering the potential application of the process (no harmful by-products generation), in order to realize practically effective HCHO oxidation.

657 5.1 Reaction temperature and time

Reaction temperature is critical in HCHO oxidation, and generally the efficiency of HCHO 658 659 oxidation over catalyst improves with increasing temperature even though some highly active catalysts are able to achieve complete oxidation at room temperature. Most of the reported 660 transition metal based catalysts attained complete HCHO oxidation at temperatures above 100°C 661 (21,36,55,67,73,82) with very few below 100 °C (19,21,73,87,95,98). Figure 6 presents the 662 conversion/removal efficiency of HCHO as a function of reaction temperatures over different 663 catalysts. However, it does not provide a basis for comparison due to the differences in the initial 664 HCHO concentration and feed flow rate employed in various experiments. This makes it difficult 665 to evaluate the specific reaction rates (equation 1) (74) of variously reported catalysts for rational 666 667 comparison. For static systems, experiments are normally conducted at room temperature; in which case HCHO oxidation becomes a function of time in addition to other factors catalyst based 668 factors discussed earlier (section 4). Figure 7 presents HCHO removal efficiency over various 669 670 catalysts as a function of reaction time. It can be seen in Figure 7 that long reaction time is required to achieve complete HCHO removal in static system. For proper comparison, other factors such as 671 initial HCHO concentration and the mass of catalyst used should be taken into account. 672

In view of these, more investigations are required to improve the activity of transition metal based catalysts at low temperature in order to compete with noble metal based catalysts which could successfully oxidize HCHO even at room temperature (*48,101*).

specific reaction rates
$$\left(\frac{\mu mol}{m^2.min}\right) = \frac{C_{HCHO} \div 22.4 \times F \times n}{S_{BET} \times m}$$
.....(1)

Where C_{HCHO} is the inlet HCHO concentration (ppm), F is the flow rate (ml/min), η is the percentage HCHO conversion (CO₂ generation), S_{BET} catalysts specific surface area and m is the catalyst mass.

680 5.2 Relative humidity

681 Relative humidity (RH) at moderate levels influences the efficiency of HCHO oxidation, however at higher RH competitive adsorption of H₂O molecules tend to block the catalyst's active surface 682 hence impinging activity. RH also helps in replenishing the supply of surface active hydroxyl 683 groups which are consumed during HCHO oxidation, via the reaction of H₂O vapor and surface 684 active oxygen (O_2^-, O_1^-) (73,102). Wang et al. (18) noted that in the presence of 50% RH, H₂O 685 vapor react with surface active oxygen to generate enough surface hydroxyl groups to sustain 686 HCHO oxidation reaction at RT. H₂O vapor also helps in carbonates desorption from the catalyst's 687 surface via competitive adsorption, thereby enabling recovery of catalytic activity (73). Wang et 688 al. (73) indicated that RH in the range of 33% to 65% enhanced HCHO removal on birnessite 689 structured manganese at room temperature. However at higher RH (92%), catalytic activity was 690 impacted owing to competitive adsorption of H₂O molecules on the catalyst's surface. Similar 691 692 effects were observed with MnO₂/PET (18): at 0% RH significant deactivation was observed due to the formation of formate species on the catalyst's surface; however at higher RH (50%), the 693 catalyst remained active and stable while activity significantly dropped when the RH was further 694

raised to 80% due to competitive adsorption of H₂O molecules. This conforms to the findings ofSidheswaran et al. (97).

697 5.3 Feed flow rate and feed concentration

698 The significance of feed (reactants) flow rate in terms of gas hourly space velocity (GHSV) on catalytic oxidation of HCHO was also reported in the literature. GHSV can improve the efficiency 699 700 of a fixed bed reaction system in two ways; at higher velocity, it will help to a certain extent in 701 enhancing external mass transfer while at a lower velocity it increases the residence time of reactants in the reactor (52). It was however reported that the efficiency of HCHO oxidation 702 703 reactions decreases with increasing space velocity (39,69,70,101) and vice-versa. In view of these, it could be insinuated that external mass transfer has little or no limitation on the efficiency 704 of HCHO oxidation reaction, although an experimental evaluation on external mass transfer effect 705 would be required. 706

In addition, HCHO feed concentration is another significant parameter for HCHO oxidation 707 708 process. HCHO oxidation efficiency tends to decrease with an increase in the feed concentration. Chuang et al. (28) demonstrated that an increase in HCHO feed concentration led to a decrease in 709 conversion at the same reaction temperature. Likewise, Tang et al. (48) reported that catalytic 710 activity decreased with rise in HCHO feed concentration from 30 to 580 ppm and that the catalysts 711 were more active and stable at lower concentration. Correspondingly, Li et al. (98) pointed out that 712 MnO₂/AC completely lost its activity when exposed to 5 mg/m³ HCHO feed concentration in 32 713 hrs, compared to its sustained HCHO oxidation efficiency of up to 70% for 80 hrs under a HCHO 714 feed concentration of 0.5 mg/m^3 . 715

716 6. Reaction mechanism of catalytic oxidation of HCHO

Understanding the reaction mechanism of HCHO oxidation over transition metal based catalysts 717 is vital for the development of catalysts with high efficiency, low cost and good stability. Different 718 719 catalysts for HCHO oxidation may exhibit varying reaction mechanisms because different intermediate species could be formed with various active oxygen species and surface active sites 720 721 (4). Oxidation reaction of HCHO over some transition metals based catalysts were shown to conform to Mars-van Krevelen mechanism (18,19,97). In this mechanism, reaction proceeds via a 722 two-stage redox reaction with an assumption of constant surface oxygen. The catalyst's surface is 723 724 first oxidized by gas phase molecular oxygen to form surface adsorbed oxygen species which subsequently reduces the pollutant (52). 725

Sekine (2) proposed that HCHO oxidation on metal oxides catalysts proceeds first by adsorption on the catalyst's surface followed by decomposition through the formation of formate intermediates on the surface, then the intermediates are subsequently decomposed to H₂O and CO₂ as presented in equations (2-5) where g and a indicate gaseous phase and adsorbed species, respectively.

Tang et al. (86) indicated the oxidation of HCHO over MnO_x -CeO₂ catalyst was attained through the effective transfer of oxygen species from CeO₂ oxygen reservoir to MnO_2 active state. This synergy is achieved through the effective activation of feed molecular oxygen and the transfer of the activated oxygen to replace the released active oxygen species from MnO_2 which participated in HCHO oxidation through a series of redox cycles involving Mn^{4+}/Mn^{3+} and Ce^{4+}/Ce^{3+} (48,86) as shown in Figure 8.

741 Wang et al. (73) proposed three-step mechanism for HCHO oxidation on birnessite structured manganese catalyst at room temperature as presented in Figure 9. The reaction first proceeds 742 743 through H-bond between HCHO molecules and birnessite bonded H2O molecules. The adsorbed molecules are then oxidized to formate and carbonates by structural hydroxyl species. 744 Subsequently, the consumed hydroxyl ions are replaced through the reaction between surface 745 746 active oxygen and H₂O molecules. An oxygen vacancy is then formed on the site of the consumed surface active oxygen species, which further acts as a site for molecular oxygen activation to active 747 species for continuous reaction (74). Wang et al. (68) reported that only hydrocarbonate species 748 were observed on the surface of 3D-Co-Mn catalysts with no obvious formation of formate species, 749 probably due to the fast conversion rate of formate to hydrocarbonates on the catalyst's surface as 750 suggested by the authors. The hydrocarbonate species were totally decomposed at 70°C as reported 751 by Shi et al. (95) for Co-Mn. 752

Shi et al. (95) indicated that the mechanism of HCHO oxidation over Co-Mn oxides proceeds via 753 the formation of DOM and formate species as intermediates. During adsorption process, HCHO is 754 immediately converted to DOM as no HCHO molecules were observed on the catalyst surface at 755 room temperature. The active DOM species react with surface adsorbed oxygen species to form 756 formate species and part of them are further oxidized to hydrocarbonates. At higher temperature 757 (50°C), further oxidation of formate and degradation of hydrocarbonate species occur. At the 758 complete oxidation temperature of HCHO (75°C), only adsorbed H2O molecules were observed, 759 indicating complete oxidation of all the intermediate species. Hence, formate oxidation and 760

hydrocarbonate decomposition are considered as the rate-limiting steps for HCHO oxidation overCo-Mn oxides catalyst.

A reaction mechanism for 2D ordered mesoporous Co₃O₄, Au/Co₃O₄ and Au/Co₃O₄-CeO₂ at room 763 temperature had been proposed by Ma et al. (51) as depicted in Figure 10. In this mechanism, 764 formate species are first generated by the nucleophilic attack on the C-H in HCHO by surface 765 active oxygen on the (110) facet of Co_3O_4 (with Co^{3+} as the active state). The formate species are 766 further oxidized to bicarbonates by surface active oxygen species. Carbonic acid species are then 767 generated by the reaction of the bicarbonates and H⁺, which are subsequently decomposed to CO₂. 768 769 In contrast, Fan et al. (67) indicated that hydroxyl groups were responsible for the immediate oxidation of HCHO on the surface of Co₃O₄ and that DOM and formate species were observed as 770 the intermediates during the adsorption of HCHO on the catalyst's surface as previously reported 771 (95). However, the behavior of the catalysts at elevated temperature in the presence of O₂ was 772 influenced by the existence of K⁺ ions on the catalyst's surface. In the presence of K⁺ ions, DOM 773 species generated during the adsorption process are converted to formate species and formate 774 oxidation to bicarbonates becomes the key reaction step. It was proposed that hydroxyl groups 775 were generated from surface hydrolysis of K₂CO₃ and got consumed during the reaction. The 776 consumed hydroxyl groups were replenished by the H₂O molecules generated thus K₂CO₃ surface 777 hydrolysis was sustained and the reaction was accelerated. On the other hand, in the absence of 778 K₂CO₃ at elevated temperature (80°C) only few DOM species could be converted to formate 779 780 species due to the lack of hydroxyl groups on the surface of Co₃O₄, thus formate decomposition to carbonates and bicarbonates became the key reaction step at \geq 80°C. Here it could be observed 781 that the existence of K⁺ on the surface of Co₃O₄ changed the reaction path from formate 782 783 decomposition to formate oxidation and continued supply of surface hydroxyl group necessary for

the oxidation reaction was sustained by K_2CO_3 surface hydrolysis thus promoting HCHO oxidation. Similar promotion effects of K⁺ was observed for Ag/Co₃O₄ (46), Na⁺ for ceramic honeycombs (103) and Na⁺ for Pd/TiO₂ (59).

787 Lu et al. (87) proposed a reaction pathway for HCHO over graphene-MnO₂ hybrid catalysts. In this mechanism, HCHO is first oxidized to form formate intermediates while molecular oxygen is 788 activated and transferred to active Mn sites through Mn⁴⁺/Mn³⁺ redox cycle. Graphene being an 789 electrical conductor reduces electron transfer resistance and enhances the rate of charge transfer 790 between Mn⁴⁺ and Mn³⁺, thereby improving the overall efficiency of the process. Thereafter, due 791 792 to the abundant amount of surface hydroxyl groups on the hybrid catalyst, formate species were directly converted to CO₂ and H₂O, while the consumed hydroxyl ions were regenerated by the 793 produced H₂O molecules as reported by other authors (67). 794

The reaction mechanism of HCHO catalytic oxidation is intricate and varies with the type and 795 amount of surface active oxygen or hydroxyl species present. But generally, formate species are 796 797 key intermediates present on almost all transition metal based catalysts. Other species such as DOM and hydrocarbonates are also reported as intermediates for HCHO oxidation. However, the 798 reported pathways for HCHO decomposition varied in the literature. In some mechanisms DOM 799 species are first generated followed by their conversion to formate intermediates, which are 800 subsequently oxidized to hydrocarbonates and finally CO₂ and H₂O molecules are generated by 801 hydrocarbonate decomposition. In other mechanisms, no DOM species are observed and formates 802 are directly oxidized to CO₂ and H₂O especially in the presence of high amount of surface hydroxyl 803 groups (67,87). The reaction mechanism of transition metal based catalysts for HCHO oxidation 804 is minimally reported in the literature. This prompts the need for further elaborate investigation 805 806 into the mechanism on various active transition metals especially as it relates to the surface properties of these catalysts, and the specific role of hydroxyl and oxygen species and the promotional effects of alkali metals (Na^+ , K^+) on the overall efficiency of the reaction. This will provide more information and offer the basis for the design of yet more active and cost-effective catalysts for practical application.

811 **7. Conclusions and outlook**

Formaldehyde is one of the most harmful indoor air pollutants as it has adverse effects on human health due to its toxicity and carcinogenicity. Techniques such as adsorption, photo-catalytic oxidation and catalytic oxidation have been used in HCHO removal. Adsorption removal is limited by the adsorption capacity of the adsorbents deployed and the hazard of HCHO desorption during regeneration while photo-catalytic oxidation produces hazardous by-products. Catalytic oxidation offsets these drawbacks and can achieve complete conversion of HCHO into CO₂ and H₂O molecules without the formation of hazardous by-products.

Noble metal based catalysts exhibit excellent low and even room temperature HCHO oxidation 819 820 capabilities but their practical application is restricted by their cost and availability. Recent investigations focus on the use of relatively cheap and abundant transition metals oxides and 821 improving their low temperature performance. More attention has mainly been focused on the 822 improvement and exploitation of various structures and morphologies of manganese oxide 823 catalysts due to its high activity yet with little attention on other transition metal catalysts. As such, 824 more investigation needs to be done on the improvement of morphological structures, textural and 825 redox properties of these materials. 826

For practical air purification application in air purifiers or building HVAC systems, substratematerials with low air resistance and pressure drop are required to immobilize catalysts to avoid

dust contamination and catalysts leaching into treated air stream. However, only a few substrate materials have been reported in the literature and hence the need for further evaluation of more materials to understand their interactions with the catalysts and effect on activity for pragmatic air purification process. In addition to the development of these materials, they need to be tested under real indoor environmental conditions in air purifiers to evaluate the effects of system conditions such as face velocity, dust particle contamination and particle leaching on the efficiency and stability of the catalysts.

No single property is decisive for catalytic performance of transition metal oxides for complete 836 837 oxidation of HCHO but a combination of several factors which include specific surface area, metal oxidation state, adsorbed and surface oxygen species, structure and morphology of the catalyst. 838 Improvement of the morphological structures of transition metal based catalysts relative to their 839 conventional bulk counterparts enhances their catalytic activity through improvement of specific 840 surface area, mesoporosity and exposure of surface active metals and oxygen species. The tunnel 841 structure of catalysts enhances the rate of adsorption of HCHO onto the active sites for 842 decomposition and the rate of product desorption from the sites to make them available for the 843 next molecules. This is particularly promoted when the catalyst's effective diameter is close to the 844 dynamic diameter of HCHO (55). Oxidation state of metals in the catalyst is also vital for 845 determining activity and higher oxidation states seem to be more active; Mn⁴⁺ in the case of 846 manganese and Co^{3+} in the case of cobalt were identified to be the most active states for HCHO 847 activation. Several methods have been proposed for improving the oxidation sate of metal oxides 848 which include modified synthesis, controlling calcination temperature and the relative molar ratio 849 of component elements, to maintain solid solution stability in the case of composite catalysts (86). 850

The use of hard template materials has been proven effective in providing the plane structure for improved exposure of more surface active sites and enhanced access to these sites.

853 Active surface oxygen and hydroxyl species directly participate in oxidizing HCHO molecules 854 and all other intermediate species into CO₂ and H₂O. The higher the concentration of these active materials on the catalyst's surface the better its HCHO oxidative activity. The availability of 855 856 surface active oxygen species and the rate of molecular oxygen activation could be enhanced by 857 the formation of surface vacancy using dopants and formation of solid solution composites with oxygen carrier materials. Recently reported works (80,81) indicated that creating oxygen surface 858 859 vacancy using dopants increases the activation rate of molecular oxygen to more active and reactive oxygen species thereby providing more abundant oxygen species for HCHO oxidation. 860 However not much has been reported on surface vacancy creation and surface modification using 861 dopants on transition metal based catalyst for HCHO oxidation, thereby prompting the need for 862 further exploration. The use of oxygen carrier materials for composite formation on the other hand 863 improves the transfer rate of active oxygen from the rich oxygen sources of the carrier to the active 864 metal sites in a complete redox cycle. The formation of solid solution within the solubility limit of 865 the composite is critical to achieving synergistic effects either through oxygen transfer or oxygen 866 867 vacancy creation, above which phase segregation occurs and synergy is lost. Most of the reported composites in the literature for HCHO oxidation focus on the use of CeO₂ as composite material 868 869 with little or no attention to other oxygen carriers.

Furthermore, enhancing the electrical conductivity of catalyst through the formation of hybrids or composites with electrically conducting materials such as graphene (*87*) was shown to greatly improve catalytic activity at low temperature. The hybrid system exposed more Mn⁴⁺ active sites, enhanced charge transport during Mn redox cycle and offered higher content of surface hydroxyl

species, which eases HCHO oxidation and improves the catalyst's activity at low temperature. 874 However, very few studies have been reported for catalyst's electrical conductivity enhancement 875 using conductor materials for HCHO oxidation. Hence further investigation into the influence of 876 dopants and other conductor materials on the activity of transition metals based catalysts for 877 improved low temperature HCHO oxidation will be desirable. In addition to catalytic properties, 878 879 reaction parameters such as reaction temperature, relative humidity, space velocity and HCHO feed concentration also affect the overall efficiency of HCHO oxidation. Relative humidity at 880 moderate levels helps in enhancing the availability of surface active hydroxyl groups via the 881 reaction of H₂O vapor and surface active oxygen but becomes detrimental at higher levels. 882

The reaction mechanism of HCHO is complex and depends on the type of catalysts, the oxidation 883 state and amount of active metals and the nature of surface active oxygen (adsorbed and or lattice) 884 present. Although further research is required to ascertain the respective role of each in HCHO 885 oxidation. Only a few mechanisms were reported for transition metal catalysts, as such further 886 investigation is required into the reaction mechanism especially as it relates to the catalyst's surface 887 properties and the role of surface active oxygen or hydroxyl species. But generally, formate 888 species, DOM and hydrocarbonates are the identified intermediates for HCHO oxidation on 889 890 transition metal based catalysts. However, the reported pathways for HCHO decomposition varied in the literature. In some mechanisms DOM species are first generated followed by DOM 891 conversion to formate intermediates, which are subsequently oxidized to hydrocarbonates and 892 finally CO₂ and H₂O molecules are generated by hydrocarbonate decomposition. In other 893 mechanisms, no DOM species are observed and formates are directly oxidized to CO2 and H2O 894 molecules without the formation of hydrocarbonates especially in the presence of high amount of 895 surface hydroxyl groups. 896

897	Acronyms	
898	1D	One Dimensional
899	2D	Two Dimensional
900	3D	Three Dimensional
901	3DOM	Three Dimensional Ordered Macroporous
902	AC	Activated Carbon
903	СТАВ	Cetyltrimethylammonium Bromide
904	DOM	Dioxymethylene
905	GHSV	Gas Hourly Space Velocity
906	НСНО	Formaldehyde
907	HVAC	Heating Ventilation and Air Conditioning
908	IARC	International Agency for Research on Cancer
909	KIT-6	Korea Advanced Institute of Science and Technology-6
910	K-OMS-2	Potassium-Octahedral Molecular Seive-2
911	PET	Polyethylene Terephthalate
912	R	Specific Reaction Rate
913	Redox	Reduction-Oxidation
914	RH	Relative humidity
915	SBA-15	Santa Barbara Amorphous-15
916	TOF	Turnover Frequency
917	VOC	Volatile Organic Compounds
918	WHO	World Health Organization

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927 References

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963

- 928 (1) Maddalena, R.; Russell, M.; Sullivan, D.P.; Apte, M.G. Formaldehyde and Other Volatile
 929 Organic Chemical Emissions in Four FEMA Temporary Housing Units. *Environ. Sci. Technol.*930 2009, 43 (15), 5626-5632.
- 932 (2) Sekine, Y. Oxidative decomposition of formaldehyde by metal oxides at room temperature.
 933 *Atmos. Environ.* 2002, *36* (35), 5543-5547.
 934
- 935 (3) Salthammer, T.; Mentese, S.; Marutzky, R. Formaldehyde in the Indoor Environment. *Chem. Rev.*936 2010, *110* (4), 2536-2572.
- 938 (4) Bai, B.; Qiao, Q.; Li, J.; Hao, J. Progress in research on catalysts for catalytic oxidation of formaldehyde. *Chin. J. Catal.* 2016, *37* (1), 102-122.
- 941 (5) Sekine, Y.; Nishimura, A. Removal of formaldehyde from indoor air by passive type air-cleaning materials. *Atmos. Environ.* 2001, *35* (11), 2001-2007.
 943
- 944 (6) International Agency for Cancer Research (IACR) working group on the identification of
 945 Carcinogenic Risks to Humans. *Formaldehyde*, 2-Butoxyethanol and 1-tert-Butoxypropan-2-ol.
 946 WHO; Lyon France, 2006; volume 88.
- 948 (7) World Health Organization. *WHO guidelines for indoor air quality: selected pollutants*; WHO:
 949 Copenhegan, 2010.

(8) Le, Y.; Guo, D.; Cheng, B.; Yu, J. Bio-template-assisted synthesis of hierarchically hollow SiO 2 microtubes and their enhanced formaldehyde adsorption performance. *Appl. Surf. Sci.* 2013, 274, 110-116.

- 955 (9) Chen, D.; Qu, Z.; Sun, Y.; Wang, Y. Adsorption–desorption behavior of gaseous formaldehyde
 956 on different porous Al2O3 materials. *Colloids Surf.*, A 2014, 441 433-440.
 957
- Ma, C.; Li, X.; Zhu, T. Removal of low-concentration formaldehyde in air by adsorption on activated carbon modified by hexamethylene diamine. *Carbon* 2011, 49 (8), 2873-2875.
- 961 (11) Rezaee, A.; Rangkooy, H.; Jonidi-Jafari, A.; Khavanin, A. Surface modification of bone char for removal of formaldehyde from air. *Appl. Surf. Sci.* 2013, 286, 235-239.
- (12) Lu, Y.; Wang, D.; Ma, C.; Yang, H. The effect of activated carbon adsorption on the photocatalytic removal of formaldehyde. *Build. Environ.* 2010, 45 (3), 615-621.
- 967 (13) Tong, H.; Zhao, L.; Li, D.; Zhang, X. N, Fe and WO3 modified TiO₂ for degradation of formaldehyde. *J. Alloys Compd.* 2011, *509* (22), 6408-6413.
 969
- 970 (14) Fu, P.; Zhang, P.; Li, J. Photocatalytic degradation of low concentration formaldehyde and
 971 simultaneous elimination of ozone by-product using palladium modified TiO 2 films under UV
 972 254+ 185nm irradiation. *Appl. Catal.*, *B* 2011, 105 (1), 220-228.
 973
- (15) Liang, W.; Li, J.; Jin, Y. Photo-catalytic degradation of gaseous formaldehyde by TiO 2/UV,
 Ag/TiO 2/UV and Ce/TiO 2/UV. *Build. Environ.* 2012, *51*, 345.-350

976 977 978	(16)	Ghosh, J.P.; Sui, R.; Langford, C.H.; Achari, G.; Berlinguette, C.P. A comparison of several nanoscale photocatalysts in the degradation of a common pollutant using LEDs and conventional UV light. <i>Water Res.</i> 2009 , <i>43</i> (18), 4499-4506.
979 980 981 982	(17)	Shie, J.L.; Lee, C.H.; Chiou, C.S.; Chang, C.T.; Chang, C.C.; Chang, C.Y. Photodegradation kinetics of formaldehyde using light sources of UVA, UVC and UVLED in the presence of composed silver titanium oxide photocatalyst. <i>J. Hazard. Mater.</i> 2008 , <i>155</i> (1), 164-172.
983 984 985 986	(18)	Wang, J.; Yunus, R.; Li, J.; Li, P.; Zhang, P.; Kim, J. In situ synthesis of manganese oxides on polyester fiber for formaldehyde decomposition at room temperature. <i>Appl. Surf. Sci.</i> 2015 , <i>357</i> , 787-794.
987 988 989 990	(19)	Zhang, J.; Li, Y.; Wang, L.; Zhang, C.; He, H. Catalytic oxidation of formaldehyde over manganese oxides with different crystal structures. <i>Catal. Sci. Technol.</i> 2015 , <i>5</i> (4), 2305-2313.
991 992 993	(20)	Luengas, A.; Barona, A.; Hort, C.; Gallastegui, G.; Platel, V.; Elias, A. A review of indoor air treatment technologies. <i>Rev. Environ. Sci. Bio/Technol.</i> 2015 , <i>14</i> (3), 499-522.
993 994 995 996 997	(21)	Chen, H.; He, J.; Zhang, C.; He, H. Self-assembly of novel mesoporous manganese oxide nanostructures and their application in oxidative decomposition of formaldehyde. <i>J. Phys. Chem. C</i> 2007 , <i>111</i> (49), 18033-18038.
998 999	(22)	Zhang, C.; He, H. A comparative study of TiO 2 supported noble metal catalysts for the oxidation of formaldehyde at room temperature. <i>Catal. Today</i> 2007 , <i>126</i> (3), 345-350.
1000 1001 1002 1003	(23)	Zhang, C.; Liu, F.; Zhai, Y.; Ariga, H.; Yi, N.; Liu, Y.; Asakura, K.; Flytzani - Stephanopoulos, M.; He, H. Alkali Metal Promoted Pt/TiO2 Opens a More Efficient Pathway to Formaldehyde Oxidation at Ambient Temperatures. <i>Angew. Chem., Int. Ed.</i> 2012, <i>51</i> (38), 9628-9632.
1004 1005 1006	(24)	Huang, H.; Leung, D.Y. Complete elimination of indoor formaldehyde over supported Pt catalysts with extremely low Pt content at ambient temperature. <i>J. Catal.</i> 2011 , <i>280</i> (1), 60-67.
1007 1008 1009	(25)	Kim, S.S.; Park, K.H.; Hong, S.C. A study on HCHO oxidation characteristics at room temperature using a Pt/TiO 2 catalyst. <i>Appl. Catal., A</i> 2011, <i>398</i> (1), 96-103.
1010 1011 1012 1013	(26)	Yan, Z.; Xu, Z.; Yu, J.; Jaroniec, M. Enhanced formaldehyde oxidation on CeO 2/AlOOH- supported Pt catalyst at room temperature. <i>Appl. Catal., B</i> 2016 , <i>199</i> , 458-465.
1014 1015 1016 1017	(27)	Nie, L.; Jiaguo, Y.; Xinyang, L.; Bei, C.; Gang, L.; Mietek, J. Enhanced performance of NaOH-modified Pt/TiO2 toward room temperature selective oxidation of formaldehyde. <i>Environ. Sci. Technol.</i> 2013 , <i>47</i> (6), 2777-2783.
1018 1019	(28)	Chuang, K.T.; Zhou, B.; Tong, S. Kinetics and Mechanism of Catalytic Oxidation of Formaldehyde over Hydrophobic Catalysts. <i>Ind. Eng. Chem. Res.</i> 1994 , <i>33</i> (7), 1680-1686.
1020 1021 1022 1023 1024	(29)	An, N.; Zhang, W.; Yuan, X.; Pan, B.; Liu, G.; Jia, M.; Yan, W.; Zhang, W. Catalytic oxidation of formaldehyde over different silica supported platinum catalysts. <i>Chem. Eng. J. (Amsterdam, Neth.)</i> 2013, <i>215</i> , 1-6.

1025 1026	(30)	Peng, J.; Wang, S. Performance and characterization of supported metal catalysts for complete oxidation of formaldehyde at low temperatures. <i>Appl. Catal.</i> , <i>B</i> 2007 , <i>73</i> (3), 282.
1027 1028 1029	(31)	Park, S.J.; Bae, I.; Nam, I.S.; Cho, B.K.; Jung, S.M.; Lee, J.H. Oxidation of formaldehyde over Pd/Beta catalyst. <i>Chem. Eng. J. (Amsterdam, Neth.)</i> 2012 , <i>195</i> , 392-402.
1030 1031 1032 1033	(32)	de la Peña O'Shea, V. A.; Alvarex-Galvan, M.C; Fierro, J.L.; Arias, P. L. Influence of feed composition on the activity of Mn and PdMn/Al ₂ O ₃ catalysts for combustion of formaldehyde/methanol. <i>Appl. Catal.</i> , <i>B</i> 2005 , <i>57</i> (3), 191-199.
1034 1035 1036	(33)	Huang, H.; Leung, D.Y. Complete Oxidation of Formaldehyde at Room Temperature Using TiO ₂ Supported Metallic Pd Nanoparticles. <i>ACS Catal.</i> 2011 , <i>1</i> (4), 348-354.
1037 1038 1039 1040	(34)	Mao, C.F.; Vannice, M. A. Formaldehyde Oxidation Over Ag Catalysts. J. Catal. 1995, 154 (2), 230-244.
1040 1041 1042 1043	(35)	Qu, Z.; Shen, S.; Chen, D.; Wang, Y. Highly active Ag/SBA-15 catalyst using post-grafting method for formaldehyde oxidation. <i>J. Mol. Catal. A: Chem.</i> 2012 , <i>356</i> , 171-177.
1043 1044 1045 1046 1047	(36)	Tian, H.; He, J.; Liu, L.; Wang, D. Effects of textural parameters and noble metal loading on the catalytic activity of cryptomelane-type manganese oxides for formaldehyde oxidation. <i>Ceram. Int.</i> 2013 , <i>129</i> (1), 315-321.
1048 1049 1050	(37)	Li, H. F.; Zhang, N.; Chen, P.; Luo, M. F.; Lu, J. Q. High surface area Au/CeO ₂ catalysts for low temperature formaldehyde oxidation. <i>Appl. Catal.</i> , <i>B</i> 2011 , <i>110</i> , 279-285.
1050 1051 1052 1053	(38)	Imamura, S.; Uchihori, D.; Utani, K.; Ito, T. Oxidative decomposition of formaldehyde on silver- cerium composite oxide catalyst <i>Catal. Lett.</i> 1994, <i>24</i> (3-4), 377-384.
1055 1054 1055 1056 1057	(39)	An, N.; Yu, Q.; Liu, G.; Li, S.; Jia, M.; Zhang, W. Complete oxidation of formaldehyde at ambient temperature over supported Pt/Fe ₂ O ₃ catalysts prepared by colloid-deposition method. <i>J. Hazard. Mater.</i> 2011 , <i>186</i> (2), 1392-1397.
1057 1058 1059 1060	(40)	Shen, Y.; Yang, X.; Wang, Y.; Zhang, Y.; Zhu, H.; Gao, L.; Jia, M. The states of gold species in CeO ₂ supported gold catalyst for formaldehyde oxidation. <i>Appl. Catal.</i> , <i>B</i> 2008 , <i>79</i> (2), 142-148.
1060 1061 1062 1063 1064	(41)	Yu, X.; He, J.; Wang, D.; Hu, Y.; Tian, H.; He, Z. Facile Controlled Synthesis of Pt/MnO ₂ Nanostructured Catalysts and Their Catalytic Performance for Oxidative Decomposition of Formaldehyde. <i>J. Phys. Chem. C</i> 2011 , <i>116</i> (1), 851-860.
1065 1066 1067	(42)	Hu, P.; Amghouz, Z.; Huang, Z.; Xu, F.; Chen, Y.; Tang, X. Surface-Confined Atomic Silver Centers Catalyzing Formaldehyde Oxidation. <i>Environ. Sci. Technol.</i> 2015 , <i>49</i> (4), 2384-2390.
1068 1069 1070 1071	(43)	Bai, B.; Qiao, Q.; Arandiyan, H.; Li, J.; Hao, J. Three-Dimensional Ordered Mesoporous MnO ₂ -Supported Ag Nanoparticles for Catalytic Removal of Formaldehyde. <i>Environ. Sci. Technol.</i> 2016 , <i>50</i> (5), 2635-2640.
1071 1072 1073 1074	(44)	Zhang, J.; Jin, Y.; Li, C.; Shen, Y.; Han, L.; Hu, Z.; Di, X.; Liu, Z. Creation of three- dimensionally ordered macroporous Au/CeO ₂ catalysts with controlled pore sizes and their enhanced catalytic performance for formaldehyde oxidation. <i>Appl. Catal.</i> , <i>B</i> 2009 , <i>91</i> (1), 11-20.

1075 1076 1077	(45)	Wang, R.; Li, J. OMS-2 Catalysts For Formaldehyde Oxidation: Effects of Ce And Pt on Structure And Performance of the Catalysts. <i>Catal. Lett.</i> 2009 , <i>131</i> (3-4), 500-505.
1078 1079	(46)	Bai, B.; Li, J. Positive Effects of K+ Ions on Three-Dimensional Mesoporous Ag/Co ₃ O ₄ Catalyst For HCHO Oxidation. <i>ACS Catal.</i> 2014 , <i>4</i> (8), 2753-2762.
1080 1081 1082	(47)	Ma, L.; Wang, D.; Li, J.; Bai, B.; Fu, L.; Li, Y. Ag/CeO ₂ nanospheres: Efficient catalysts for formaldehyde oxidation. <i>Appl. Catal.</i> , <i>B</i> 2014 , <i>148</i> , 36-43.
1083 1084 1085	(48)	Tang, X.; Chen, J.; Huang, X.; Xu, Y.; Shen, W. Pt/MnO _x –CeO ₂ catalysts for the complete oxidation of formaldehyde at ambient temperature. <i>Appl. Catal.</i> , <i>B</i> 2008 , <i>81</i> (1), 115-121.
1086 1087 1088 1089	(49)	Tang, X.; Chen, J.; Li, Y.; Li, Y.; Xu, Y.; Shen, W. Complete oxidation of formaldehyde over Ag/MnO _x -CeO ₂ catalysts. <i>Chem. Eng. J. (Amsterdam, Neth.)</i> 2006, <i>118</i> (1), 119-125.
1090 1091 1092	(50)	Liu, B.; Liu, Y.; Li, C.; Hu, W.; Jing, P.; Wang, Q.; Zhang, J. Three-dimensionally ordered macroporous Au/CeO ₂ -Co ₃ O ₄ catalysts with nanoporous walls for enhanced catalytic oxidation of formaldehyde. <i>Appl. Catal., B</i> 2012 , <i>127</i> , 47-58.
1093 1094 1095 1096 1097 1098	(51)	Ma, C.; Wang, D.; Xue, W.; Dou, B.; Wang, H.; Hao, Z. Investigation of Formaldehyde Oxidation over Co ₃ O ₄ –CeO ₂ and Au/Co ₃ O ₄ –CeO ₂ Catalysts at Room Temperature: Effective Removal and Determination of Reaction Mechanism. <i>Environ. Sci. Technol.</i> 2011 , <i>45</i> (8), 3628-3634.
1099 1100 1101	(52)	Pei, J.; Zhang, J.S. Critical review of catalytic oxidization and chemisorption methods for indoor formaldehyde removal. <i>HVACR Res.</i> 2011 , <i>17</i> (4), 476-503.
1102 1103 1104	(53)	Quiroz, J.; Giraudon, J.M.; Gervasini, A.; Dujardin, C.; Lancelot, C.; Trentesaux, M.; Lamonier, J.F. Total Oxidation of Formaldehyde over MnO _x -CeO ₂ Catalysts: The Effect of Acid Treatment. <i>ACS Catal.</i> 2015 , <i>5</i> (4), 2260-2269.
1105 1106 1107 1108	(54)	Tian, H.; He, J.; Liu, L.; Wang, D.; Hao, Z.; Ma, C. Highly active manganese oxide catalysts for low-temperature oxidation of formaldehyde. <i>Microporous Mesoporous Mater.</i> 2012 , <i>151</i> , 397-402.
1109 1110 1111 1112	(55)	Chen, T.; Dou, H.; Li, X.; Tang, X.; Li, J.; Hao, J. Tunnel structure effect of manganese oxides in complete oxidation of formaldehyde. <i>Microporous Mesoporous Mater.</i> 2009 , <i>122</i> (1), 270-274.
1112 1113 1114 1115 1116	(56)	Tian, H.; He, J.; Zhang, X.; Zhou, L.; Wang, D. Facile synthesis of porous manganese oxide K-OMS-2 materials and their catalytic activity for formaldehyde oxidation. <i>Microporous Mesoporous Mater.</i> 2011 , <i>138</i> (1), 118-122.
1110 1117 1118 1119	(57)	Tasbihi, M.; Bendyna, J. K.; Notten, P. H. A Short Review on Photocatalytic Degradation of Formaldehyde. <i>J. Nanosci. Nanotechnol.</i> 2015 , <i>15</i> (9), 6386-6396.
1120 1121	(58)	Chen, B.B.; Shi, C.; Crocker, M.; Wang, Y.; Zhu, A.M. Catalytic removal of formaldehyde at room temperature over supported gold catalysts. <i>Appl. Catal.</i> , <i>B</i> 2013 , <i>132</i> , 245-255.
1122 1123 1124	(59)	Zhang, C.; Li, Y.; Wang, Y.; He, H. Sodium-Promoted Pd/TiO ₂ for Catalytic Oxidation of Formaldehyde at Ambient Temperature. <i>Environ. Sci. Technol.</i> 2014 , <i>48</i> (10), 5816-5822.

1125 1126	(60)	Xu, Z.; Yu, J.; Jaroniec, M. Efficient catalytic removal of formaldehyde at room temperature using AlOOH nanoflakes with deposited Pt. <i>Appl. Catal.</i> , <i>B</i> 2015 , <i>163</i> , 306-312.
1127 1128 1129	(61)	Li, D.; Yang, G.; Li, P.; Wang, J.; Zhang, P. Promotion of Formaldehyde Oxidation over Ag Catalyst by Fe Doped MnO _x Support at Room Temperature. <i>Catal. Today</i> 2016 , <i>277</i> , 257-265.
1130 1131	(62)	Zhang, Y.; Shen, Y.; Yang, X.; Sheng, S.; Wang, T.; Adebajo, M. F.; Zhu, H. Gold catalysts
1132 1133		supported on the mesoporous nanoparticles composited of zirconia and silicate for oxidation of formaldehyde. <i>J. Mol. Catal. A: Chem.</i> 2010 , <i>316</i> , 100-105.
1134 1135 1136 1137	(63)	Li, C., Shen, Y.; Jia, M.; Sheng, S.; Adebajo, M. O.; Zhu, H. Catalytic combustion of formaldehyde on gold/iron-oxide catalysts. <i>Catal. Commun.</i> 2008 , <i>9</i> (3), 355-361.
1137 1138 1139 1140 1141	(64)	Xu, Q.; Lei, W.; Li, X.; Qi, X.; Yu, J.; Liu, G.; Wang, J.; Zhang, P. Efficient Removal of Formaldehyde by Nanosized Gold on Well-Defined CeO ₂ Nanorods at Room Temperature. <i>Environ. Sci. Technol.</i> 2014 , <i>48</i> (16), 9702-9708.
1142 1143 1144 1145	(65)	Chen, H.; Tang, M.; Rui, Z.; Ji, H. MnO ₂ Promoted TiO ₂ Nanotube Array Supported Pt Catalyst for Formaldehyde Oxidation with Enhanced Efficiency. <i>Ind. Eng. Chem. Res.</i> 2015 , <i>54</i> (36), 8900-8907.
1146 1147 1148	(66)	Chen, Y.; He, J.; Tian, H.; Wang, D.; Yang, Q. Enhanced formaldehyde oxidation on Pt/MnO 2 catalysts modified with alkali metal salts. <i>J. Colloid Interface Sci.</i> 2014 , <i>428</i> , 1-7.
1149 1150 1151 1152	(67)	Fan, Z.; Zhang, Z.; Fang, W.; Yao, X.; Zou, G.; Shangguan, W. Low-temperature catalytic oxidation of formaldehyde over Co_3O_4 catalysts prepared using various precipitants. <i>Chin. J. Catal.</i> 2016 , <i>37</i> (6), 947-954.
1153 1154 1155 1156	(68)	Wang, Y.; Zhu, A.; Chen, B.; Crocker, M.; Shi, C. Three-dimensional ordered mesoporous Co- Mn oxide: A highly active catalyst for "storage–oxidation" cycling for the removal of formaldehyde. <i>Catal. Commun.</i> 2013 , <i>36</i> , 52.
1157 1158 1159 1160 1161	(69)	Xia, Y.; Dai, H.; Zhang, L.; Deng, J.; He, H.; Au, C.T. Ultrasound-assisted nanocasting fabrication and excellent catalytic performance of three-dimensionally ordered mesoporous chromia for the combustion of formaldehyde, acetone and methanol. <i>Appl. Catal., B</i> 2010 , <i>100</i> (1), 229-237.
1162 1163 1164	(70)	Bai, B.; Arandiyan, H.; Li, J. Comparison of the performance for oxidation of formaldehyde on nano-Co ₃ O ₄ , 2D-Co ₃ O ₄ , and 3D-Co ₃ O ₄ catalysts. <i>Appl. Catal.</i> , <i>B</i> 2013 , <i>142</i> , 677-683.
1165 1166 1167	(71)	Dai, Z.; Yu, X.; Huang, C.; Li, M.; Su, J.; Guo, Y.; Xu, H.; Ke, Q. Nanocrystalline MnO ₂ on an activated carbon fiber for catalytic formaldehyde removal. <i>RSC Adv.</i> 2016 , <i>6</i> (99), 97022-97029.
1167 1168 1169 1170	(72)	Feng, Q.; Kanoh, H.; Ooi, K. Manganese oxide porous crystals. J. Mater. Chem. 1999, 9 (2), 319-333.
1171 1172 1173	(73)	Wang, J.; Zhang, P.; Li, J.; Jiang, C.; Yunus, R.; Kim, J. Room-Temperature Oxidation of Formaldehyde by Layered Manganese Oxide: Effect of Water. <i>Environ. Sci. Technol.</i> 2015 , <i>49</i> (20), 12372-12379.

1174 1175 1176	(74)	Wang, J.; Li, J.; Jiang, C.; Zhou, P.; Zhang, P.; Yu, J. The effect of manganese vacancy in birnessite-type MnO ₂ on room-temperature oxidation of formaldehyde in air. <i>Appl. Catal., B</i> 2017, <i>204</i> , 147-155.
1177 1178 1179 1180	(75)	Wang, J.; Zhang, G.; Zhang, P. Layered birnessite-type MnO ₂ with surface pits for enhanced catalytic formaldehyde oxidation activity. <i>J. Mater. Chem. A</i> 2017 , <i>5</i> (12), 5719.
1180 1181 1182 1183	(76)	Bai, B.; Qiao, Q.; Li, J.; Hao, J. Synthesis of three dimensional ordered mesoporous MnO2 and its catalytic performance in formaldehyde oxidation. <i>Chin. J. Catal.</i> 2016 , <i>37</i> , 27-31.
1185 1184 1185 1186 1187	(77)	Zhou, L.; Zhang, J.; He, J.; Hu, Y.; Tian, H. Control over the morphology and structure of manganese oxide by tuning reaction conditions and catalytic performance for formaldehyde oxidation. <i>Mater. Res. Bull.</i> 2011 , <i>46</i> (10), 1714-1722.
1188 1189 1190 1191	(78)	Wu, Y.; Ma, M.; Zhang, B.; Gao, Y.; Lu, W.; Guo, Y. Controlled synthesis of porous Co3O4 nanofibers by spiral electrospinning and their application for formaldehyde oxidation. <i>RSC Adv.</i> 2016 , <i>6</i> , 102127-102133.
1192 1193 1194	(79)	Zeng, L.; Li, K.; Huang, F.; Zhu, X.; Li, H. Effects of Co ₃ O ₄ nanocatalyst morphology on CO oxidation: Synthesis process map and catalytic activity. <i>Chin. J. Catal.</i> 2016 , <i>37</i> , 908-922.
1195 1196 1197 1198	(80)	Huang, Y.; Long, B.; Tang, M.; Rui, Z.; Balogun, M.S.; Tong, Y.; Ji, H. Bifunctional catalytic material: An ultrastable and high-performance surface defect CeO 2 nanosheets for formaldehyde thermal oxidation and photocatalytic oxidation. <i>Appl. Catal.</i> , <i>B</i> 2016 , <i>181</i> , 779-787.
1190 1199 1200 1201 1202	(81)	Zeng, L.; Song, W.; Li, M.; Zeng, D.; Xie, C. Catalytic oxidation of formaldehyde on surface of H TiO2/H C TiO2 without light illumination at room temperature. <i>Appl. Catal., B</i> 2014 , <i>147</i> , 490-498.
1202 1203 1204 1205 1206	(82)	Zhou, L.; He, J.; Zhang, J.; He, Z.; Hu, Y.; Zhang, C.; He, H. Facile In-Situ Synthesis of Manganese Dioxide Nanosheets on Cellulose Fibers and their Application in Oxidative Decomposition of Formaldehyde. <i>J. Phys. Chem. C</i> 2011 , <i>15</i> (34), 16873-16878.
1207 1208 1209	(83)	Imamura, S.; Shono, M.; Okamoto, N.; Hamada, A.; Ishida, S. Effect of cerium on the mobility of oxygen on manganese oxides. <i>Appl. Catal., A</i> 1996, <i>142</i> (2), 279-288.
1210 1211 1212 1213	(84)	Luo, M. F.; Ma, J. M.; Lu, J. Q.; Song, Y. P.; Wang, Y.J. High-surface area CuO–CeO ₂ catalysts prepared by a surfactant-templated method for low-temperature CO oxidation. <i>J. Catal.</i> 2007 , <i>246</i> (1), 52-59.
1214 1215 1216	(85)	Ding, Z.Y.; Li, L.; Wade, D.; Gloyna, E.F. Supercritical water oxidation of NH ₃ over a MnO ₂ /CeO ₂ catalyst <i>Ind. Eng. Chem. Res.</i> 1998 , <i>37</i> (5), 1707-1716.
1217 1218 1219 1220	(86)	Tang, X.; Li, Y.; Huang, X.; Xu, Y.; Zhu, H.; Wang, J.; Shen, W. MnO _x –CeO ₂ mixed oxide catalysts for complete oxidation of formaldehyde: effect of preparation method and calcination temperature. <i>Appl. Catal.</i> , <i>B</i> 2006 , <i>62</i> (3), 265-273.
1220 1221 1222 1223	(87)	Lu, L.; Tian, H.; He, J.; Yang, Q. Graphene–MnO ₂ Hybrid Nanostructure as a New Catalyst for Formaldehyde Oxidation. <i>J. Phys. Chem. C</i> 2016 , <i>120</i> (41), 23660-23668.

1224 1225 1226 1227 1228	(88)	Lou, Y.; Cao, XM.; Lan, J.; Wang, L.; Dai, Q.; Guo, Y.; Ma, J.; Zhao, Z.; Guo, Y.; Hu, P.; Lu, G. Ultralow-temperature CO oxidation on an $In_2O_3-Co_3O_4$ catalyst: a strategy to tune CO adsorption strength and oxygen activation simultaneously. <i>Chem. Commun. (Cambridge, U.K.)</i> 2014 , <i>50</i> (52), 6835-6838.
1228 1229 1230 1231	(89)	Wang, Z.; Pei, J.; Zhang, J. Catalytic oxidization of indoor formaldehyde at room temperature– Effect of operation conditions. <i>Build. Environ.</i> 2013 , <i>65</i> , 49-57.
1232 1233 1234 1235	(90)	Wen, Y.; Tang, X.; Li, J.; Hao, J.; Wei, L.; Tang, X. Impact of synthesis method on catalytic performance of MnOx–SnO2 for controlling formaldehyde emission. <i>Catal. Commun.</i> 2009 , <i>10</i> (8), 1157-1160.
1236 1237 1238 1239	(91)	Bai, L.; Wyrwalski, F.; Lamonier, J.F.; Khodakov, A.Y.; Monflier, E.; Ponchel, A. Effects of β- cyclodextrin introduction to zirconia supported-cobalt oxide catalysts: From molecule-ion associations to complete oxidation of formaldehyde. <i>Appl. Catal.</i> , <i>B</i> 2013 , <i>138</i> , 381-390.
1240 1241 1242 1243	(92)	Xuesong, L.; Jiqing, L.; Kun, Q.; Huang, W; Mengfei, L. A comparative study of formaldehyde and carbon monoxide complete oxidation on MnO _x -CeO ₂ catalysts. <i>J. Rare Earths</i> 2009 , <i>27</i> (3), 418-424.
1244 1245 1246 1247	(93)	Li, J.W.; Pan, K.L.; Yu, S.J.; Yan, S.Y.; Chang, M.B. Removal of formaldehyde over MnxCe1– xO ₂ catalysts: Thermal catalytic oxidation versus ozone catalytic oxidation <i>J. Environ. Sci.</i> 2014 , <i>26</i> , 2546-2553.
1247 1248 1249 1250	(94)	Zhu, L.; Wang, J.; Rong, S.; Wang, H.; Zhang, P. Cerium modified birnessite-type MnO 2 for gaseous formaldehyde oxidation at low temperature. <i>Appl. Catal., B</i> 2017, <i>211</i> , 212-221.
1250 1251 1252 1253	(95)	Shi, C.; Wang, Y.; Zhu, A.; Chen, B.; Au, C. MnxCo ₃ –xO ₄ solid solution as high-efficient catalysts for low-temperature oxidation of formaldehyde. <i>Catal. Commun.</i> 2012 , <i>28</i> , 18-22.
1253 1254 1255 1256	(96)	Lu, S.; Li, K.; Huang, F.; Chen, C.; Sun, B. Efficient MnO _x -Co ₃ O ₄ -CeO ₂ catalysts for formaldehyde elimination. <i>Appl. Surf. Sci.</i> 2017 , <i>400</i> , 277-282.
1257 1258 1259 1260	(97)	Sidheswaran, M. A.; Destaillats, H.; Sullivan, D. P.; Larsen, J.; Fisk, W. J. Quantitative room-temperature mineralization of airborne formaldehyde using manganese oxide catalysts. <i>Appl. Catal., B</i> 2011, <i>107</i> (1), 34-41.
1261 1262 1263 1264	(98)	Li, J., Zhang, P., Wang, J. and Wang, M. Birnessite-Type Manganese Oxide on Granular Activated Carbon for Formaldehyde Removal at Room Temperature. <i>J. Phys. Chem. C</i> 2016 , <i>120</i> (42), 24121-24129.
1265 1266 1267	(99)	Bai, B.; Li, J; Hao, J. 1D-MnO ₂ , 2D-MnO ₂ and 3D-MnO ₂ for low-temperature oxidation of ethanol. <i>Appl. Catal.</i> , <i>B</i> 2015 , <i>164</i> , 241-250.
1267 1268 1269 1270 1271	(100)	Sun, Y.; Zhang, W.; Li, D.; Gao, L.; Hou, C.; Zhang, Y.; Liu, Y. Facile synthesis of MnO ₂ /rGO/Ni composite foam with excellent pseudocapacitive behavior for supercapacitors. <i>J. Alloys Compd.</i> 2015 , <i>649</i> , 579-584.
1272 1273 1274	(101)	Zhang, C.; He, H.; Tanaka, K.I. Catalytic performance and mechanism of a Pt/TiO2 catalyst for the oxidation of formaldehyde at room temperature. <i>Appl. Catal.</i> , <i>B</i> 2006 , <i>65</i> (1), 37-43.

- 1275 (102) Chen, B.B.; Zhu, X. B.; Crocker, M.; Wang, Y.; Shi, C. FeO x-supported gold catalysts for catalytic removal of formaldehyde at room temperature. *Appl. Catal.*, *B* **2014**, *154*, 73-81.
- 1277
- 1278 (103) Yu, J.; Li, X.; Xu, Z.; Xiao, W. NaOH-Modified Ceramic Honeycomb with Enhanced
 1279 Formaldehyde Adsorption and Removal Performance. *Environ. Sci. Technol.* 2013, 47 (17),
 1280 9928-9933.

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Catalyst	Preparation method	Reaction conditions	HCHO conversion/removal	Reference
1 wt% Pd/TiO ₂	Deposition precipitation-reduction	500 mg catalyst, 10 ppm HCHO, air, 50% ³ RH, 120,000 mL/g·h ⁵ GHSV	² 100% at 25°C	(33)
1 wt% Au/CeO ₂	Deposition precipitation (using urea)	250 mg catalysts,80 ppm HCHO, 21%O ₂ /N ₂ , 50% ³ RH, 34,000/h ⁵ GHSV	¹ 100% at 25°C	(58)
1 wt% Au/CeO ₂	Deposition precipitation (using NaOH)	250 mg catalyst, 80 ppm HCHO, 21%O ₂ /N ₂ , 50% ³ RH, 34,000/h ⁵ GHSV	¹ 100% at 70°C	(58)
2 wt% Na-1 wt% Pd/TiO ₂	Impregnation-reduction	65 mg catalyst, 140 ppm HCHO, 20% O ₂ , 25% ³ RH, 95,000/h ⁵ GHSV	¹ 100% at 25°C	(59)
0.8 wt% Pt/AlOOH	Impregnation-reduction	100 mg catalyst, 138 ppm HCHO, air, 25% ³ RH	² 96.82% ⁴ (25°C,1h)	(60)
0.8 wt% Pt/TiO2	Impregnation-reduction	100 mg catalyst, 138 ppm HCHO, air, 25% ³ RH	² 96.82% ⁴ (25°C,1h)	(60)
Ag/CeO ₂ (nanosphere)	Hydrothermal synthesis	50 mg catalyst, 810 ppm HCHO, 20% O ₂ , and balance N ₂ , 84,000/h ⁵ GHSV	¹ 100% at 110°C	(47)
Ag/3D-Co ₃ O ₄	Deposition- precipitation	200 mg catalyst, 100 ppm HCHO, 20% O2, 30,000/h ⁵ GHSV	¹ 100% at 100°C	(46)
1.7wt% K- Ag/3D- Co3O4	Deposition- precipitation	200 mg catalyst, 100 ppm HCHO, 20% O ₂ , 30,000/h ⁵ GHSV	¹ 100% at 70°C ¹ 55% at 27°C	(46)
8.9 wt% Ag/3D- MnO ₂	Deposition- precipitation	200 mg catalyst, 500 ppm HCHO, 20% O ₂ , 60,000/h ⁵ GHSV	¹ 100% at 110°C	(43)
Ag/MnO ₂	Redox-reaction	200 mg catalyst, 230 ppm HCHO, air, static reactor	¹ 76% ⁴ (25°C,1h)	(61)
Ag/Fe _{0.1} - MnO ₂	Redox-reaction	200 mg catalyst, 230 ppm HCHO, air, static reactor	¹ 100% ⁴ (25°C, 1h)	(61)
Ag/MnO ₂	Redox-reaction	200 mg catalyst, 400 ppm HCHO, 21% O ₂ , 30,000 mL/g·h ⁵ GHSV, dynamic system	¹ 100% at 120°C	(61)
Ag/Fe _{0.1} - MnO ₂	Redox-reaction	200 mg catalyst, 400 ppm HCHO, 21% O ₂ , 30,000 mL/g·h ⁵ GHSV, dynamic system	¹ 100% at 90°C	(61)
3DOM-Au/CeO ₂	Nanocasting: Polystyrene colloidal crystal hard templates	200 mg catalyst, 0.06% HCHO, purified air, 66,000 mL/ g·h ⁵ GHSV	¹ 100% at 75°C	(44)
0.85 wt% Au/ZrO ₂ - silicate	Deposition	200 mg catalyst, 90 mg/m ³ HCHO, purified air, 52,000 ml/g·h ⁵ GHSV	¹ 100% at 157°C	(62)
3 wt% Au/CeO2	Deposition- precipitation	50 mg catalyst, 500 ppm HCHO, 20% O ₂ , 35,400/h ⁵ GHSV	¹ 92.3% at 37°C and ¹ 100% at 50°C	(37)
7.10 wt% Au/Fe ₂ O ₃	Co-precipitation	200 mg catalyst, 6.25mg/m ³ HCHO, compressed air, 54,000 ml/g·h ⁵ GHSV	¹ 100% at 80°C	(63)
1.8 wt % Au/CeO ₂	Deposition- precipitation	150 mg catalyst, 109.3 ppm HCHO, air, static reactor	² 90% (25°C, 1h)	(64)
2 wt Pt/urchin-like MnO ₂	RT redox reaction	100 mg catalyst, 460 ppm, purified air 20,000 mL/ g·h ⁵ GHSV	¹ 100% at 80°C	(41)

2 wt% Pt/cocoon-like MnO2	Redox reaction under acidic condition	100 mg catalyst, 460 ppm HCHO, purified air 20,000 mL/ g·h ⁵ GHSV	¹ 100% at 90°C	(41)
Pt/OMS-2	Redox reaction plus impregnation	200 mg catalyst, 500 ppm HCHO, 10% O ₂ , 30,000 ml/ g·h ⁵ GHSV	¹ 100% at 120°C	(45)
0.2 wt% Pt/0.5 wt% MnO ₂ /TiO ₂ (Nanotubes)	Electrochemical anodization plus impregnation	200 mg catalyst, 50 ppm HCHO, 20% O ₂ , 35% ³ RH, 30,000 ml/g·h ⁵ GHSV	² 95% at 30°C	(65)
0.1 wt% Pt/TiO ₂	Impregnation-reduction	500 mg catalyst, 10 ppm HCHO, air, 50% ³ RH, 80,000/h ⁵ GHSV	² 99.1% at 25°C	(24)
2 wt% Na-1 wt% Pt/MnO ₂ (birnessite)	Deposition- precipitation	100 mg catalyst, 200 ppm HCHO, purified air, 30,000 mL/g·h ⁵ GHSV	¹ 100% at 50°C	(66)
NaOH modified - 1 wt% Pt/TiO ₂	Impregnation-reduction	300 mg catalyst, 253 ppm HCHO, air, static reactor	² 94.07% ⁴ (25°C, 1h)	(27)
1 wt% Pt/TiO ₂	Impregnation	100 ppm HCHO, 20 vol% O ₂ , 50,000/h ⁵ GHSV	¹ 100% at 20°C	(22)
1 wt% Rh/TiO ₂	Impregnation	100 ppm HCHO, 20 vol% O ₂ , 50,000/h ⁵ GHSV	¹ 100% at 80°C	(22)
1 wt% Pd/TiO ₂	Impregnation	100 ppm HCHO, 20 vol% O ₂ , 50,000/h ⁵ GHSV	¹ 100% at 120°C	(22)
3D 3wt% Au/ CeO ₂ - Co ₃ O ₄ (2.5:1, Ce/Co)	Nanocasting: 3D-PS hard template	200 mg catalyst, 8 ppm HCHO, purified air, 15,000 mL/g·h ⁵ GHSV	¹ 100% at 36°C	(50)
2D 1wt% Au/ Co ₃ O ₄ - CeO ₂ (7:3 Co:Ce)	Nonocasting: SBA-15 hard template	100 mg catalyst, 200 ppm HCHO, 22% O ₂ , 55,000/h ⁵ GHSV	¹ 50% at 25°C	(51)
3 wt% Pt/MnO _x -CeO ₂ (M.R = 0.5)	Impregnation	200mg catalyst, 30-580 ppm HCHO, 20% O ₂ , 30,000 mL/g·h ⁵ GHSV	¹ 100% at 25°C	(48)

¹Conversion = [CO₂] _{out} / [HCHO] _{in} ²Removal = ([HCHO] _{in} - [HCHO] _{out}) / [HCHO] _{in} ³RH - Relative humidity ⁴Indicates the reaction temperature and time taken to reach stated removal/conversion in a static reactor. ⁵GHSV- Gas hourly space velocity (h⁻¹ or ml/g.h)

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Catalyst	Preparation method	Reaction conditions	HCHO conversion/removal	Reference
Birnessite	Hydrothermal synthesis using surfactant (SDS)	100 mg catalyst, 460 ppm HCHO, purified air, 30,000 ml/g·h ⁵ GHSV	¹ 100% at 100°C	(54)
Birnessite	Redox Reaction of KMnO4 and Ammonium oxalate	100 mg catalyst , 40 ppm HCHO, air, 80% ³ RH, 120,000 ml/g·h,	¹ 90% at 96°C	(74)
Birnessite	Surface modification of birnesite with nitric acid and Tetraammonium hydroxide	100 mg catalyst, 200 ppm HCHO, air, , 45% ³ RH, 120,000 ml/g·h ⁵ GHSV	¹ 100% at 100°C	(75)
Todorokite	Na-Birnesite followed by MgCl ₂ hydrothermal treatment	200 mg catalyst, 400 ppm HCHO, 10.0% O ₂ , 18,000 ml/g·h ⁵ GHSV	¹ 100% at 160°C	(55)
Pyrolusite	Redox hydrothermal synthesis	200 mg catalyst, 400 ppm HCHO, 10.0% O ₂ , 18,000 ml/g·h ⁵ GHSV	¹ 100% at 180°C	(55)
K-OMS-2 nanoparticle	Soft chemistry (KMnO4 and benzyl alcohol) using surfactant (CTAB) at 25°C	100 mg catalyst, 460 ppm HCHO, purified air, 20,000 ml/g·h ⁵ GHSV	¹ 64% at 100°C.	(56)
K-OMS-2 nanorod	Soft chemistry (KMnO4 and benzyl alcohol) using surfactant (CTAB) at 100°C	100 mg catalyst, 460 ppm HCHO, purified air, 20,000 ml/g·h ⁵ GHSV	¹ 10% at 100°C	(56)
K-OMS-2 nanorods	Sol-gel synthesis at 70°C	100 mg catalyst, 460 ppm HCHO, 21% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 100% at 200°C	(36)
K-OMS-2 nanoparticles	Sol-gel synthesis at 15°C	100 mg catalyst, 460 ppm HCHO, 21% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 54% at 200°C	(36)
OMS-2	Redox reaction	200 mg catalyst, 500 ppm HCHO, 10% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 100% at 120°C	(45)
3D-MnO ₂ mesoporous	Nanocasting: KIT-6 hard template	200 mg catalyst, 400 ppm HCHO, 20% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 100% at 130°C	(76)
α -MnO ₂ nanorods	Redox-hydrothermal synthesis	200 mg catalyst, 400 ppm HCHO, 20% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 100% at 140°C	(76)
β -MnO ₂ nanorods	Redox-hydrothermal synthesis	200 mg catalyst, 400 ppm HCHO, 20% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 100% at 180°C	(76)
8.86% MnO ₂ /cellulose (nanosheet birnessite)	KMnO ₄ impregnation followed by oleic acid reduction	60 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50, 000/h ⁵ GHSV	² 100% at 140°C	(82)
Birnessite nanospheres	KMnO4-oleic acid hydrothermal reduction	50 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50,000/h ⁵ GHSV	² 100% at 140°C	(77)
Cryptomelane nanorods	KMnO ₄ -oleic acid hydrothermal reduction	50 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50,000/h ⁵ GHSV	² 95.1% at 140°C	(77)
Ramsdellite nanorods	KMnO ₄ oleic acid hydrothermal reduction	50 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50,000/h ⁵ GHSV	² 87.2% at 140°C	(77)
Monoclinic MnOOH	KMnO ₄ oleic acid hydrothermal reduction	50 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50,000/h ⁵ GHSV	² 90.1% at 140°C	(77)

nanosheets	electrodeposition of Eu onto CeO ₂	25% O ₂ , 30,000mL/g·h ⁵ GHSV 100 mg catalyst, 100 ppm	² 100% at 120°C	(80)
Co ₃ O ₄ nanofibers 4% Eu doped-CeO ₂	spiral electrospinning synthesis, calcined at 500°C Anodic	100 mg catalyst, 400 ppm HCHO, 20% O ₂ , 30,000 mL/g·h ⁵ GHSV 200 mg catalyst, 50 ppm HCHO,	¹ 100% at 98°C	(78)
2D- Co ₃ O ₄	Nanocasting: SBA-15 hard template	100 mg catalyst, 200 ppm HCHO, 22% O ₂ , 55,000/h ⁵ GHSV	¹ 20.3% at 25°C	(51)
3D-Cr ₂ O ₃	Nanocasting: KIT-6 hard template	100 mg catalyst, 500 ppm HCHO, 30,000 ml/g·h ⁵ GHSV	¹ 90% at 117°C	(69)
3D- Co ₃ O ₄	Nanocasting: KIT-6 hard template	200 mg catalyst, 100 ppm HCHO, 20% O ₂ , 30,000/h ⁵ GHSV	¹ 100% at 110°C	(46)
Nano- Co ₃ O ₄	Precipitation with Na ₂ CO ₃	200 mg catalyst, 400 ppm HCHO, 20% O ₂ , 30,000 ml/g·h ⁵ GHSV	¹ 100% at 230°C	(70)
2D- Co ₃ O ₄	Nanocasting: SBA-15 hard template	200 mg catalyst, 400 ppm HCHO, 20% O₂, 30,000 ml/g·h ⁵ GHSV	¹ 100% at 150°C	(70)
3D- Co ₃ O ₄	Nanocasting: KIT-6 hard template	200 mg catalyst, 400 ppm HCHO, 20% O₂, 30,000 ml/g·h ⁵ GHSV	¹ 100% at 130°C	(70)
Spinel Co ₃ O ₄	Precipitation with NH ₃ ·H ₂ O	100 mg catalyst, 100 ppm HCHO, 21% O ₂ , 69,000/h ⁵ GHSV	¹ 100% at 130°C	(67)
Spinel Co ₃ O ₄	Precipitation with KOH	100 mg catalyst, 100 ppm HCHO, 21% O ₂ , 69,000/h	¹ 100% at 120°C	(67)
Spinel Co ₃ O ₄	Precipitation with NH4HCO3 and 2% K2CO3 treatment	100 mg catalyst, 100 ppm HCHO, 21% O ₂ , 69,000/h ⁵ GHSV	¹ 100% at 90°C	(67)
Spinel Co ₃ O ₄	Precipitation with KHCO ₃	100 mg catalyst, 100 ppm HCHO, 21% O ₂ , 69,000/h ⁵ GHSV	¹ 100% at 90°C	(67)
δ-MnO ₂	Hydrothermal synthesis	60 mg catalyst, 170 ppm HCHO, 20% O ₂ , 25% ³ RH, 100,000 mL/g·h ⁵ GHSV	¹ 100% at 80°C	(19)
γ-MnO ₂	Hydrothermal synthesis	60 mg catalyst, 170 ppm HCHO, 20% O ₂ , 25% ³ RH, 100,000 mL/g·h ⁵ GHSV	¹ 100% at 150°C	(19)
β-MnO ₂	Hydrothermal synthesis	60 mg catalyst, 170 ppm HCHO, 20% O ₂ , 25% ³ RH, 100,000 mL/g·h ⁵ GHSV	¹ 100% at 200°C	(19)
α-MnO ₂	Hydrothermal synthesis	60 mg catalyst, 170 ppm HCHO, 20% O ₂ , 25% ³ RH, 100,000 mL/g·h ⁵ GHSV	¹ 100% at 125°C	(19)
Birnessite	Facile redox-reaction	50 mg catalyst, 200 ppm HCHO, synthetic air, static reactor	² 100% ⁴ (25°C,12h)	(73)
KxMnO ₂ mesoporous honeycomb nanospheres	High concentration KMnO4-oleic acid reduction	70 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50,000/h	² 100% at 85°C	(21)
KxMnO ₂ mesoporous hollow nanospheres	Low concentration KMnO4-oleic acid reduction	50 mg catalyst, 100 ppm HCHO, 20% O ₂ , 50,000/h ⁵ GHSV	² 100% at 80°C	(21)

¹Conversion = [CO₂] _{out} / [HCHO] _{in} ²Removal = ([HCHO] _{in} - [HCHO] _{out}) / [HCHO] _{in} ³RH - relative humidity ⁴Indicates the reaction temperature and time taken to reach such conversion in a static reactor.

⁵GHSV- Gas hourly space velocity (h⁻¹ or ml/g.h)

Table 3	
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Catalyst	Preparation method	Reaction conditions	HCHO conversion/removal	Reference
MnO _x -CeO ₂	Co-precipitation	200 mg catalyst, 580 ppm HCHO, 18% O ₂ , 21,000 mL/g·h ⁴ GHSV	¹ 100% at 100°C	(86)
MnO _x -CeO ₂	Co-precipitation	200 mg catalyst, 580 ppm HCHO, 20% O ₂ , 30,000 mL/g·h ⁴ GHSV	¹ 90% at 90°C	(48)
Mn0.5Ce0.5O2	Co-precipitation	300 mg catalyst, 33 ppm HCHO, 21% O ₂ , 10,000/h ⁴ GHSV	² 100% at 270°C	(93)
Mn0.5Ce0.5O2	Co-precipitation	300 mg catalyst, 61 ppm HCHO, 21% O ₂ /506 ppm HCHO O ₃ , 10,000/h ⁴ GHSV	² 100% at 25°C	(93)
MnO _x -CeO ₂	Co-precipitation	200mg catalyst, 580 ppm HCHO, 20% O ₂ , 30,000 mL/g·h ⁴ GHSV	¹ 100% at 125°C	(92)
Ce-MnO ₂	Redox reaction of KMnO4 and Ammonium oxalate with Cerium nitrate	100 mg catalyst, 190 ppm HCHO, air, 90 L/g·h ⁴ GHSV	¹ 100% at 100°C	(94)
3D-CeO ₂ - Co ₃ O ₄	Nanocasting: 3D-PS hard template	200 mg catalyst, 8 ppm HCHO, purified air, 15,000 mL/g·h ⁴ GHSV	¹ 100% at 155°C	(50)
Co-Mn	Co-precipitation	150 mg catalysts, 80 ppm HCHO, 21% O ₂ , 50% ³ RH, 60,000/h ⁴ GHSV	¹ 100% at 75°C	(95)
3D-Co-Mn	Nanocasting: KIT 6 Hard template	250 mg catalyst, 80 ppm HCHO, 21% O ₂ , 50% ³ RH, 36,000/h ⁴ GHSV	¹ 100% at 70°C	(68)
MnO _x -Co ₃ O ₄ - CeO ₂	Sol-gel synthesis	50 mg catalyst, 200 ppm HCHO, 21% O ₂ , 36,000 ml/g·h ⁴ GHSV	¹ 100% at 100°C	(96)
MnO _x -SnO ₂	Redox co-precipitation	200 mg catalyst, 400 ppm HCHO, 10% O ₂ , 30,000 ml/g·h ⁴ GHSV	¹ 100% at 180°C	(90)
MnO _x -SnO ₂	co-precipitation	200 mg catalyst, 400 ppm HCHO, 10% O ₂ , 30,000 ml/g·h ⁴ GHSV	¹ 100% at 220°C	(90)
Graphene-MnO ₂ hybrid	Graphene treatment with KMnO4 (redox)	100 mg catalyst, 100 ppm HCHO, purified air, 30,000 mL/g·h ⁴ GHSV	¹ 100% at 65°C	(87)

¹ Conversion = $[CO_2]_{out}$ / $[HCHO]_{in}$ ² Removal = ($[HCHO]_{in}$ - $[HCHO]_{out}$ / $[HCHO]_{in}$ ³ RH - relative humidity ⁴GHSV- Gas hourly space velocity (h⁻¹ or ml/g.h)

Table 4

Catalyst	Preparation method	Reaction conditions	HCHO removal	Reference
δ-MnO ₂ /PET 2	Redox-precipitation	500 mg catalyst, 0.6 mg/m ³ HCHO, purified air, 50% RH, 17,000/h ² GHSV	94% at 25°C	(18)
8.86 wt% MnO ₂ /cellulose fiber 2	KMnO ₄ impregnation of cellulose fiber followed by oleic acid treatment	5 mg catalyst, 100 ppm HCHO HCHO, 20 vol % O ₂ , 50 000/h ² GHSV	99.1% at 140°C	(82)
Birnessite/AC 2	KMnO4 reduction of	200 mg catalyst, 400 ppm HCHO, synthetic air	100% ¹ (25°C,7h)	(98)
δ-MnO ₂ /AC 2	Redox co-precipitation	100 mg catalyst, 150 ppm HCHO, pure air	100% ¹ (25°C,9h)	(71)

¹ Indicates the reaction temperature and time taken to reach such conversion in a static reactor.

²GHSV- Gas hourly space velocity (h⁻¹ or ml/g.h)

Figure Captions

Figure 1. (a) Effect of birnessite calcination temperature on HCHO oxidation activity at room temperature. (b) HCHO conversion and CO₂ generation (*73*). "Reproduced with permission from Wang, J.; Zhang, P.; Li, J.; Jiang, C.; Yunus, R.; Kim, J. Room-Temperature Oxidation of Formaldehyde by Layered Manganese Oxide: Effect of Water. Environ. Sci. Technol. 2015, 49 (20), 12372-12379. Copyright (2015) American Chemical Society".

Figure 2. Effect of HCHO concentration on Pt/MnO_x-CeO₂ activity at room temperature.
Reprinted (*48*). (Reproduced with permission from Tang, X.; Chen, J.; Huang, X.; Xu, Y.; Shen,
W. Pt/MnO_x-CeO₂ catalysts for the complete oxidation of formaldehyde at ambient temperature.
Appl. Catal., B 2008, 81 (1), 115-121. Copyright, 2008 Elsevier).

Figure 3. Concentration variation of HCHO and CO₂ formation in a static experiment: (a) over birnessite-MnO₂/AC (*98*) "Reproduced with permission from Li, J., Zhang, P., Wang, J. and Wang, M. Birnessite-Type Manganese Oxide on Granular Activated Carbon for Formaldehyde Removal at Room Temperature. J. Phys. Chem. C 2016, 120 (42), 24121-24129. Copyright (2016) American Chemical Society" and (b) over δ-MnO₂/AC (*71*). (Reproduced from Dai, Z.; Yu, X.; Huang, C.; Li, M.; Su, J.; Guo, Y.; Xu, H.; Ke, Q. Nanocrystalline MnO₂ on an activated carbon fiber for catalytic formaldehyde removal. RSC Adv. 2016, 6 (99), 97022-97029, with permission from The Royal Society of Chemistry).

Figure 4. Effect of reaction condition on the evolution of morphology and structure manganese based catalysts (77). (Reproduced with permission from Zhou, L.; Zhang, J.; He, J.; Hu, Y.; Tian, H. Control over the morphology and structure of manganese oxide by tuning reaction conditions

and catalytic performance for formaldehyde oxidation. Mater. Res. Bull. 2011, 46 (10), 1714-1722. Copyright, 2011 Elsevier).

Figure 5. Effect of tunnel structure on the complete oxidation of HCHO (*55*). (Reproduced with permission from Chen, T.; Dou, H.; Li, X.; Tang, X.; Li, J.; Hao, J. Tunnel structure effect of manganese oxides in complete oxidation of formaldehyde. Microporous Mesoporous Mater. 2009, 122 (1), 270-274. Copyright, 2009 Elsevier).

Figure 6. Formaldehyde conversion over transition metal catalysts as a function of reaction temperature.

Figure 7. Formaldehyde conversion over transition metal catalysts as a function of reaction time in a static system.

Figure 8. Synergistic Oxygen transfer mechanism in MnO_x-CeO₂ composites (*48*). (Reproduced with permission from Tang, X.; Chen, J.; Huang, X.; Xu, Y.; Shen, W. Pt/MnO_x-CeO₂ catalysts for the complete oxidation of formaldehyde at ambient temperature. Appl. Catal., B 2008, 81 (1), 115-121. Copyright, 2008 Elsevier).

Figure 9. Mechanism of HCHO oxidation on birnessite at RT (*73*). "Reproduced with permission from Wang, J.; Zhang, P.; Li, J.; Jiang, C.; Yunus, R.; Kim, J. Room-Temperature Oxidation of Formaldehyde by Layered Manganese Oxide: Effect of Water. Environ. Sci. Technol. 2015, 49 (20), 12372-12379. Copyright (2015) American Chemical Society".

Figure 10. Mechanism of HCHO oxidation over mesoporous Co₃O₄, Au/Co₃O₄, and Au/Co₃O₄-CeO₂ catalysts at RT (*51*). "Reproduced with permission from Ma, C.; Wang, D.; Xue, W.; Dou, B.; Wang, H.; Hao, Z. Investigation of Formaldehyde Oxidation over Co₃O₄-CeO₂ and Au/Co₃O₄-CeO₂ Catalysts at Room Temperature: Effective Removal and Determination of Reaction Mechanism. Environ. Sci. Technol. 2011, 45 (8), 3628-3634. Copyright (2011) American Chemical Society".

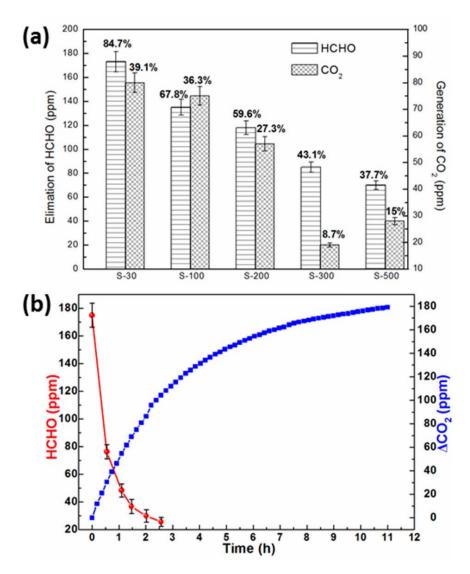


Figure 1

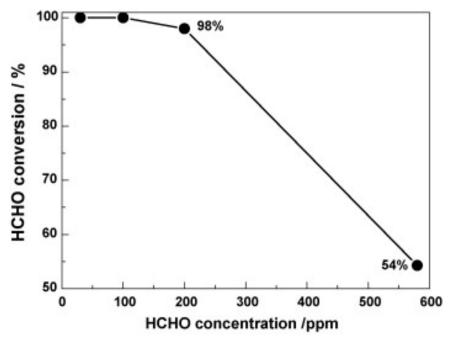


Figure 2

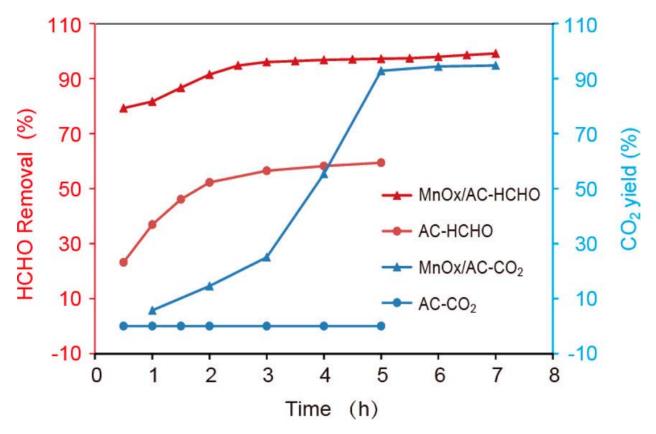


Figure 3(a)

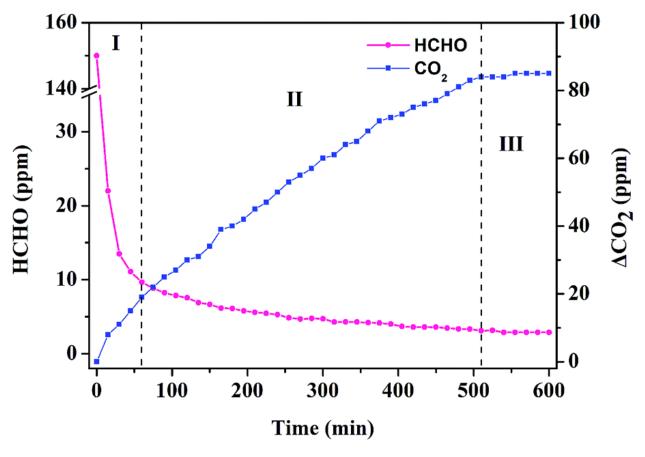


Figure 3(b)

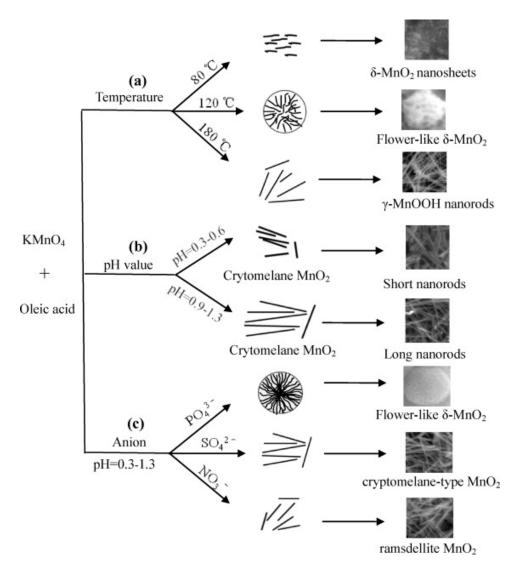
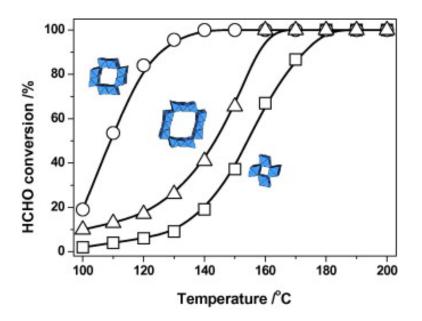


Figure 4





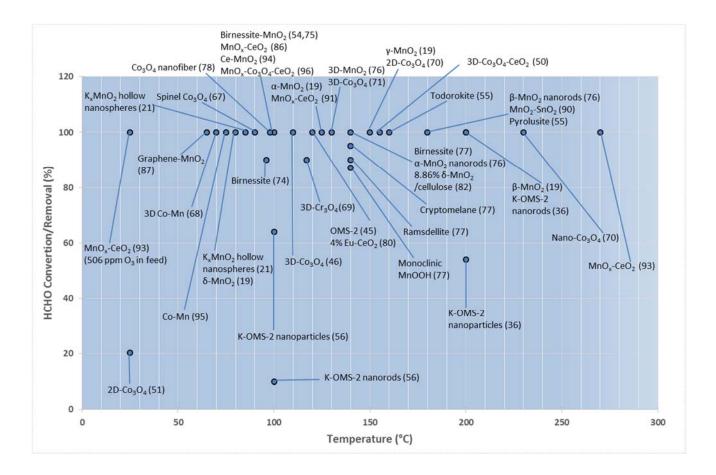


Figure 6

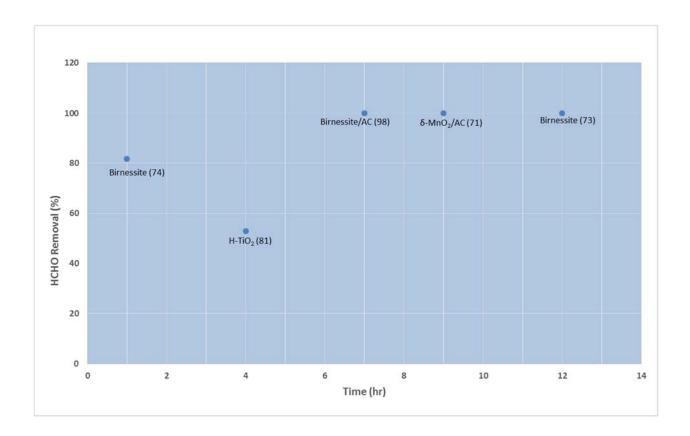


Figure 7

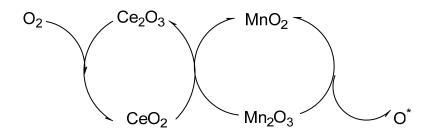


Figure 8

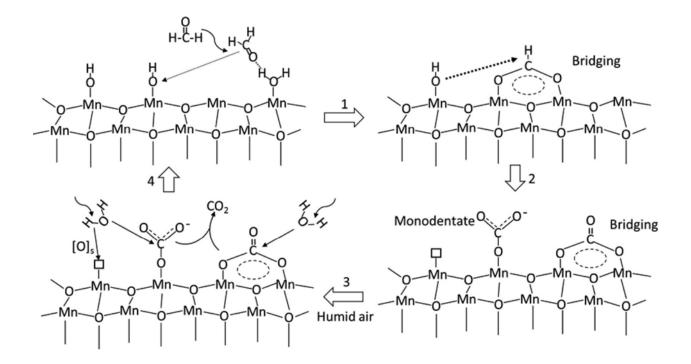


Figure 9

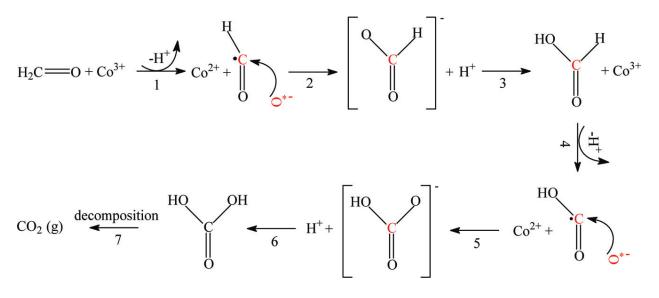


Figure 10